



Live benthic foraminiferal faunas from the French Mediterranean Coast: towards a new biotic index of environmental quality

Christine Barras, Frans Jorissen, C. Labrune, B. Andral, P. Boissery

► To cite this version:

Christine Barras, Frans Jorissen, C. Labrune, B. Andral, P. Boissery. Live benthic foraminiferal faunas from the French Mediterranean Coast: towards a new biotic index of environmental quality. *Ecological Indicators*, 2014, 36, pp.719 - 743. 10.1016/j.ecolind.2013.09.028 . hal-03278074

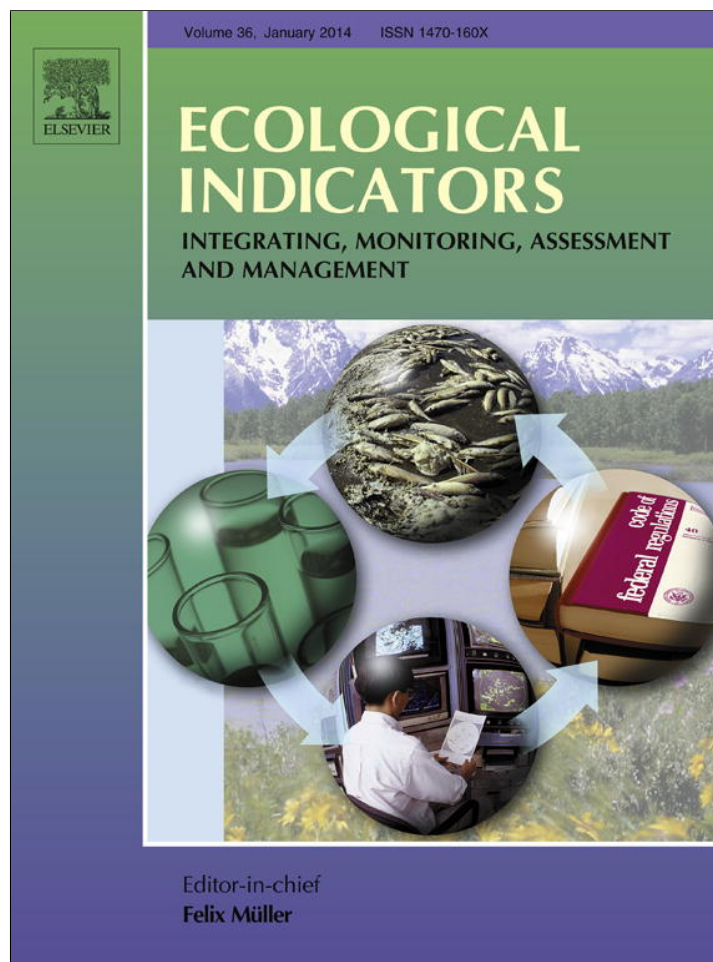
HAL Id: hal-03278074

<https://univ-angers.hal.science/hal-03278074>

Submitted on 5 Jul 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



This article appeared in a journal published by Elsevier. The attached copy is furnished to the author for internal non-commercial research and education use, including for instruction at the authors institution and sharing with colleagues.

Other uses, including reproduction and distribution, or selling or licensing copies, or posting to personal, institutional or third party websites are prohibited.

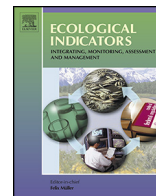
In most cases authors are permitted to post their version of the article (e.g. in Word or Tex form) to their personal website or institutional repository. Authors requiring further information regarding Elsevier's archiving and manuscript policies are encouraged to visit:

<http://www.elsevier.com/authorsrights>



Contents lists available at ScienceDirect

Ecological Indicators

journal homepage: www.elsevier.com/locate/ecolind

Live benthic foraminiferal faunas from the French Mediterranean Coast: Towards a new biotic index of environmental quality

Christine Barras^{a,*}, Frans J. Jorissen^a, Céline Labrune^{b,c}, Bruno Andral^d, Pierre Boissery^e^a UMR CNRS 6112 LPG-BIAF, Bio-Indicateurs Actuels et Fossiles, Angers University, 2 Boulevard Lavoisier, 49045 Angers Cedex, France^b UPMC, Université Paris 06, Observatoire Océanologique, 66650 Banyuls/Mer, France^c UMR CNRS 8222 LECOB, Observatoire Océanologique, 66650 Banyuls/Mer, France^d IFREMER, Laboratoire Environnement Ressources Provence-Azur-Corse, Centre Méditerranée – Zone Portuaire de Brégailhon – BP 330, 83507 La Seyne-sur-Mer Cedex, France^e Agence de l'Eau Rhône-Méditerranée et Corse, 2-4 allée de Lodz, 69363 Lyon Cedex 07, France

ARTICLE INFO

Article history:

Received 10 April 2013

Received in revised form

17 September 2013

Accepted 21 September 2013

Keywords:

Coastal Mediterranean Sea

Foraminiferal faunas

Indicative species

Tolerant species

Biotic index

Water Framework Directive

ABSTRACT

In this study, living (Rose Bengal stained) foraminiferal faunas from 31 stations along the entire French Mediterranean Sea coast except Corsica have been analysed. In the context of the Water Framework Directive, the aim was to develop a biotic index to evaluate the benthic ecosystem quality. Therefore, different faunal parameters (diversity indices, wall structure proportion, and indicative species groups) have been tested to determine their relevance as indicators of environmental conditions. The best results are obtained with a biotic index based on the relative proportion of stress-tolerant taxa. For ecosystem quality evaluation, it is essential to distinguish between natural and anthropogenic eutrophication phenomena. In order to do so, we applied a correction on our biotic index, using the expected percentage of stress-tolerant species in natural environments, in function of sediment grain size (percentage <63 µm). Finally, a comparison of the different faunal parameters calculated for two different sediment intervals (0–1 and 0–4 cm) indicates clearly that the analysis of the uppermost centimetre of the sediment is sufficient to obtain relevant information needed for bio-monitoring purposes.

© 2013 Elsevier Ltd. All rights reserved.

1. Introduction

Due to their strategic location at the interface of marine and terrestrial areas, coastal ecosystems have been impacted by human activities since the advent of human societies. Anthropogenic impact in coastal marine ecosystems has multiple origins, such as urban sewage, industrial and agricultural activities or fisheries, and results in environmental problems, such as eutrophication, oxygen deficiency, chemical pollution or physical disturbance. Awareness of recent changes in ecological conditions in many coastal seas has fostered a need to assess increasing anthropogenic pressures and their consequences on sediment and water quality, and to suggest measures to reverse this trend. In this context, the European commission implemented the Water Framework Directive (WFD, Directive 2000/60/EC) with the aim to obtain (or to maintain) a “good status” for all the European waters by 2015. The WFD defines the ecological status as the quality of the structure and functioning of ecosystems and is assessed using different planktonic and benthic indicators (e.g. phytoplankton, macro-algae, angiosperms, fish

faunas and soft substrate benthic invertebrate fauna) (Devlin et al., 2007).

The study of the benthic macrofauna is the traditional tool for benthic ecological quality assessment and bio-monitoring studies, since macrofauna responds in a predictable way to anthropogenic and natural stress (Pearson and Rosenberg, 1978). Environmental managers need an easily interpretable ecological quality status based on quantitative data. Numerous biotic indicator methods were developed for macrofauna (see review in Diaz et al., 2004) based either on diversity indices (e.g. Shannon index, Pielou, 1975) or on indices based on the relative proportions of faunal groups with different ecological characteristics. Some of the latter methods are based on groups with different feeding strategies (e.g. ITI, Word, 1979), whereas others distinguish several classes of pollution-sensitive versus opportunistic, pollution-tolerant, species (e.g. AMBI, Borja et al., 2000; BENTIX, Simbora and Zenetos, 2002; BOPA, Gomez Gesteira and Dauvin, 2000; Dauvin and Ruellet, 2007; BQI, Rosenberg et al., 2004).

More recently, benthic foraminiferal faunas have been increasingly used as bio-indicators of anthropogenic pollution. Initially, foraminifera were mainly studied in fossil records for biostratigraphic and paleoenvironmental purposes. The interest for the living organisms greatly expanded when researchers started to

* Corresponding author. Tel.: +33 2 41 73 50 02.

E-mail address: christine.barras@univ-angers.fr (C. Barras).

study their ecology in 1960s. Because of their short life cycle (3 months to 2 years, Murray, 1991), these organisms are able to respond rapidly to environmental changes, with a change in diversity and in species composition. Such rapid adaptive responses have been observed in response to changes in the quantity and quality of organic supplies (e.g. Altenbach and Sarnthein, 1989; Corliss and Emerson, 1990; Corliss, 1991; Herguera and Berger, 1991; Rathburn and Corliss, 1994; Jorissen et al., 1995, 1998; De Rijk et al., 2000; Licari et al., 2003), in oxygen conditions (e.g. Sen Gupta and Machain-Castillo, 1993; Gooday, 1994; Jorissen et al., 1995; Gooday et al., 2000), pH (e.g. Murray, 1989), salinity and temperature (e.g. Murray, 2006). Moreover, foraminifera are ubiquitous in marine environments, inhabiting transitional to abyssal areas and tropical to polar latitudes (review in Murray, 2006). Foraminifera are abundant in marine sediments, even in deep-sea environments where they commonly represent more than 50% of the total biomass (Gooday et al., 1992). The high number of individuals sampled with a little quantity of sediment assures the robustness of data analysis and limits the impact of sampling on the seafloor. Furthermore, foraminiferal taxonomy is easy compared to the identification of macrofauna, since only a single biological group is considered, instead of several phyla. Although foraminifera represent only a part of the trophic niches and guilds, the ecological characteristics of the different species are different enough to obtain reliable information about the environmental conditions, as it has been shown in a wide range of papers on benthic foraminiferal ecology (e.g. Gooday and Rathburn, 1999; Jorissen et al., 2007; Murray, 2006). Finally, the main advantage of foraminifera is the conservation of a large part of their tests (shells) in the sediment after their death. The study of dead faunas at different depths in the sediment can give important information about the natural conditions which existed before a site became polluted. This is especially useful in case of the absence of an environmental baseline study (Alve, 1995). Comparison of living faunas and pre-impact faunas can also yield essential information about individual species ecological strategies. For example, opportunistic species which have colonised the area, or sensitive species which disappear from the area after the onset of pollution, can easily be recognised. As such, the comparison of live and dead faunas can ensure that lists with species ecological characteristics correctly translate the behaviour of the various species at the study site.

All these advantages make foraminifera an innovative and very interesting tool for bio-monitoring studies of anthropogenic impact (reviews in Alve, 1995; Nigam et al., 2006; Frontalini and Coccioni, 2011). The first studies using foraminifera as indicators of environmental quality appeared in 1960s (Resig, 1960; Watkins, 1961; Bandy et al., 1964, 1965; Seiglie, 1968, 1971; Clark, 1971). Today, numerous studies use foraminifera as bio-indicators of different types of pollution such as eutrophication (e.g. Platon et al., 2005; Mojtabid et al., 2008; Hyams-Kaphzan et al., 2009), heavy metals (e.g. Alve, 1991; Armynot du Chatelet et al., 2004; Frontalini and Coccioni, 2008; Bergamin et al., 2009; Cherchi et al., 2009; Coccioni et al., 2009; Frontalini et al., 2009; Romano et al., 2009; Vilela et al., 2011), urban sewage (e.g. Burone et al., 2006; Teodoro et al., 2010), oil drilling activities (e.g. Durrieu et al., 2006; Mojtabid et al., 2006; Duchemin et al., 2008; Jorissen et al., 2009; Denoyelle et al., 2010), oil spills (e.g. Morvan et al., 2004) or aquaculture (e.g. Bouchet et al., 2007). However, no standardised protocols for sampling and sampling treatment have been defined until recently (Schönfeld et al., 2012) so that direct comparison of the various studies is very difficult, if not impossible. However, a careful observation of the faunal patterns described in these studies allows identifying different types of species behaviour in response to pollution. Just as for macrofauna, some studies tried to develop biotic indices, either based on faunal diversity (e.g. Bouchet et al., 2012) or on the

proportion of indicative species (e.g. Mojtabid et al., 2006; Jorissen et al., 2009).

In the present study, we analyse living (Rose Bengal stained) foraminiferal faunas from the French Mediterranean Sea coast (except Corsica) in the context of the WFD, with the aim to evaluate the ecosystem quality. The study area represents more than 1000 km of coastal zone for which the presence of anthropogenic stress parameters is badly known. There are no point sources of pollution close to sampling stations, and there are no well-defined reference stations exempt of any anthropogenic impact either. Therefore, we first analysed the various faunal parameters (faunal density, diversity and faunal composition) that could be used for the evaluation of the environmental quality. Next, we tried to take into account the natural variability of the system in order to distinguish between the impact of this natural variability and a putative anthropogenic impact. Unfortunately, our study was performed prior to the establishment of a standardised sampling and sampling treatment protocol by the FOBIMO group (Schönfeld et al., 2012), and therefore does not follow all recommendations made in this paper. However, by comparing the faunal data for the 0–1 cm and 0–4 cm sediment levels, we tested the possibility to restrict faunal analyses to the topmost centimetre, as recommended by Schönfeld et al. (2012). By studying only the topmost centimetre, the time needed for picking the foraminifera would be largely reduced, making the method better adapted for cost-efficient bio-monitoring studies. This study represents the first crucial step for the development of a new biotic index based on benthic foraminiferal faunas. In order to be used routinely in future surveys, the presented index will need to be tested in cases with a strong pollution gradient and in other geographic areas. Since it is the first large scale study of living foraminiferal faunas along the entire French Mediterranean coast (except Corsica), the results of the present study can also serve as a global inventory and a baseline for future studies.

2. Materials and methods

2.1. Regional setting of the study area

The Mediterranean Sea is generally considered as a semi-enclosed oligotrophic basin. Low salinity surface water from the Atlantic Ocean enters the Mediterranean Sea through the Strait of Gibraltar and creates the Liguro-Provençal Current (LPC) which flows along the French Mediterranean coast, from Italy to Spain through the Gulf of Lion (Millot and Taupier-Letage, 2005; Pairaud and Desmare, 2011). The LPC can develop small scale gyres, depending of the background stratification or external forcing that can influence the shelf circulation.

There is a clear difference in the continental shelf characteristics along the Mediterranean French coast. The continental shelf in front of the Provence Côte d'Azur region is relatively narrow, less than 1 mile wide (Pairaud and Desmare, 2011). East of Toulon, an area with rocky sea floor is interrupted by several small embayments containing more fine-grained sediments, such as the Bay of Villefranche, between Villefranche and Nice. On the western side of the French Mediterranean coast, the continental shelf of the Gulf of Lion is wide (up to 40 miles; Bassetti et al., 2006) and consists of a large crescent shaped area incised by sub-marine canyons (Berné and Gorini, 2005). The bottom sediment distribution displays a mid-shelf mud belt and the inner and outer shelf regions with mixed sandy to muddy deposits (Aloisi et al., 1973).

The Gulf of Lions is also strongly influenced by the Rhône River input (Raimbault and Durrieu de Madron, 2003). With a mean annual discharge of 1700 m³/s (Thill et al., 2001), the Rhône is one of the main sources of freshwater and organic carbon for the Mediterranean Sea (Pont, 1997; Sempéré et al., 2000). The Rhône River has

a mean sediment discharge of about $9.9 \pm 6.4 \times 10^9$ kg/yr (Sempéré et al., 2000; Pont et al., 2002), accounting for 80% of the riverine input to the Gulf of Lions (Durrieu de Madron et al., 2000). The Rhône prodelta is characterised by silty muds with high organic carbon content (1–2%; Durrieu de Madron et al., 2000) and very high sediment accumulation rates. Also smaller coastal rivers (e.g. the Têt and Hérault Rivers) can significantly contribute to the sediment budget in this area (Kim et al., 2006).

Finally, this part of the Mediterranean Sea is characterised by endemic *Posidonia* seagrass meadows. In our study area, *Posidonia* meadows are located in front of Banyuls-sur-Mer (Blanc-Vernet, 1969, 1984; Vénec-Peyré and Le Calvez, 1981, 1988; Vénec-Peyré, 1984) and form a continuous band from the east side of the Rhône prodelta to the Italian frontier (Boudouresque et al., 2006).

2.2. Sampling strategy

From March 26th to April 9th 2009, 31 stations were sampled for the study of benthic foraminiferal faunas along the French Mediterranean coast on board of the research vessel “Europe” (Fig. 1, Appendix A). The location of the stations was chosen according to the WFD criteria, i.e. within one mile from the coastline and at least one station per water body (i.e. a coherent geographic area based on physical (e.g. hydrodynamic, sedimentological) criteria influencing biological activities).

2.3. Foraminiferal sampling methods

Surface sediment was sampled using a Reineck box corer, which was subsampled with plexiglass cores (diameter 7.1 cm). Only station Cerbère could be sampled with an interface corer (Gemax twin corer, core diameter 8.8 cm).

On board, cores were sliced horizontally, every half centimetre from the surface to 2 cm depth, every centimetre between 2 and 6 cm depth, and every two centimetres from 6 to 10 cm depth. Sometimes, cores were too short to sample until 10 cm depth. For stations Gruissan, Lavandou and Faraman, it was not possible to

take a core in the Reineck box (e.g. because of the presence of many pebbles), and the first centimetre of the surface was sampled with a spoon. In this case, after homogenisation, 50 cm³ of sediment was subsampled for foraminiferal analyses.

After sampling, sediments were stored in plastic bottles filled with a mixture of ethanol (95%) and Rose Bengal stain (1 g/l). Rose Bengal is commonly used to obtain a rapid overview of the living faunas. It stains the cytoplasm of foraminifera alive at the time of sampling (Walton, 1952), or which died in a recent past (weeks to months, Bernhard, 1988; Corliss and Emerson, 1990), and in which the non-degraded proteins are still stainable. Ethanol allows preserving stained cellular tissues for a prolonged period of time. Samples were gently shaken to obtain a homogeneous mixture and were transported to the laboratory for further processing.

2.4. Foraminiferal analyses

In the laboratory, sediment samples treated with Rose Bengal were sieved through 150 and, if necessary, 500 µm mesh screens. For our study, only the >150 µm or 150–500 µm fraction was analysed, depending on the station. The >500 µm size fraction was removed when the sediment contained large quantities of vegetal detritus, shell fragments or coarse sand, which complicated foraminiferal picking. The >500 µm fraction was checked on some occasions and no living foraminifera were found. We consider therefore that in our study area, the results obtained for the 150–500 µm size fraction are comparable with those of the >150 µm fraction. Also Bouchet et al. (2012) observed that the number of individuals >500 µm in their samples from the Norwegian Skagerrak coast was minimal. Unfortunately, our foraminiferal analyses were performed prior to the establishment of the methodological recommendations of the FOBIMO group (Schönfeld et al., 2012). The main differences between our methodology and the one described by the FOBIMO group is the use of the >150 µm instead of the >125 µm size fraction, and the absence of replicate cores, which could not be sampled due to time constraints.

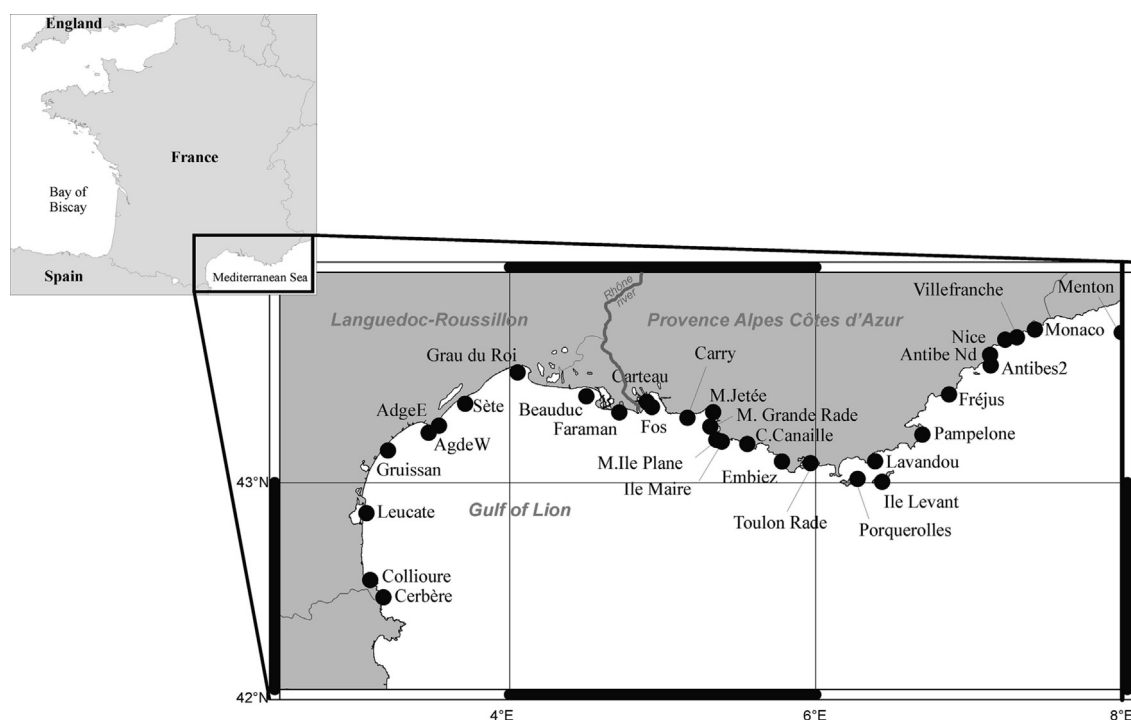


Fig. 1. Localisation of the sampling stations along the French Mediterranean coast.

Rose Bengal stained foraminifera were wet-picked in 50% ethanol under a binocular microscope (Leica MZ95). Only specimens showing a clear pink colour (or red, depending on the species) in all but the last chambers were considered as living fauna. If necessary, opaque porcelaneous and agglutinated specimens were broken to check for the presence of protoplasm. Next, foraminifera were arranged on micropaleontological slides, identified on species level using taxonomic handbooks, and counted.

In order to study the vertical distribution (and microhabitats) of living foraminifera in the sediment, 14 stations have been analysed until at least 4 cm depth in sediment (station Toulon Grande Rade has only been sampled until 3 cm) and a maximum of 10 cm depth. The faunal parameters from the 0–4 cm sediment interval have been compared to those obtained for the 0–1 cm interval, to determine whether the study of deeper sediment intervals (time-consuming and therefore more expensive) yields important complementary information. An important aim of the present study was to determine whether the study of the 0–1 cm sediment interval is sufficient to describe the quality of the benthic ecosystem, if so, supporting one of the recommendations of the FOBIMO group (Schönfeld et al., 2012).

2.5. Foraminiferal parameters

For each station and studied sediment interval (i.e. 0–1 cm or 0–4 cm), we calculated the following faunal parameters: (1) total foraminiferal density (standardised for a 50 cm² sediment surface), (2) specific richness, and (3) the respective proportion of the three principal foraminiferal groups (perforate, porcelaneous and agglutinated foraminifera). To describe the diversity of the foraminiferal faunas, we used the Shannon–Wiener *H* index (Hayek and Buzas, 1997) and the equitability *J* index (Pielou, 1966) which are defined by the following equations:

$$H = - \sum \left(\left(\frac{n_i}{N} \right) \times \ln \left(\frac{n_i}{N} \right) \right) \quad \text{and} \quad J = \frac{H}{\ln(S)}$$

where n_i is the number of individuals of species i , N is the total number of individuals, and S is the total number of species at the considered station. The Shannon–Wiener index links the number of species to the assemblage density whereas the equitability index focuses particularly on the distribution of individual densities between the different species (it distinguishes between samples with comparable densities for all species or samples with a dominance of one or a few species).

Because foraminiferal abundances are very different between stations, we also calculated (using PAST software, Hammer and Harper, 2005) the expected number of species from a sub-sample of 50 individuals taken from the population of all the individuals (ES₅₀). The concept of expected number of species (ES) was first introduced by Sanders (1968) but its computation was modified by Hurlbert (1971). It is computed as:

$$ES_{50} = 1 - \sum_{i=1}^s \frac{(N - N_i)!(N - 50)!}{(N - N_i - 50)!N!}$$

where N is the total abundance of individuals at the considered station, N_i is the abundance of the i th species at the considered station, and s is the number of species at the considered station. ES₅₀ was not calculated when absolute density was lower than 50 individuals.

After testing the data for normal distribution (Shapiro–Wilk test adapted to small size samples, $n < 50$), we used parametric (Student test) or non-parametric (Wilcoxon test) statistical analyses for paired samples in order to compare the data obtained for 0–1 cm and 0–4 cm sediment intervals. Differences were considered significant when $p < 0.05$.

We also studied the vertical distribution of the various taxa in the first centimetres of sediment. Foraminiferal microhabitats are controlled by physical, chemical and biological processes (Corliss, 1985; Buzas et al., 1993; Jorissen et al., 1995). The microhabitat concept allows a better understanding of the food and oxygen needs of each species. Therefore we calculated the Average Living Depth (ALD_x) for the total fauna of the core as follows (Jorissen et al., 1995):

$$ALD_x = \sum_{i=1,x} \frac{(n_i \times D_i)}{N}$$

in which ALD_x is the average living depth (in cm) of the fauna in a core of x centimetres depth; n_i is the number of specimens in the sediment interval i ; D_i is the midpoint of the sediment interval i (in cm); and N is the total number of individuals for all levels.

2.6. Environmental parameters

Pore water oxygen profiles were measured on board under *in situ* temperature conditions using a cathode-type mini-electrode (100 or 500 μm tips, Unisense®) (Revsbech, 1983; Helder and Bakker, 1985; Revsbech and Jørgensen, 1986) for Reineck cores with a well-preserved sediment water interface with overlying bottom waters. These analyses were generally duplicated and allowed to determine the maximum oxygen penetration depth (OPD) in the sediment.

During the oceanographic cruise, in addition to sediment for foraminiferal analysis, sediment was also sampled for grain size and total organic matter analyses. Grain size analysis was conducted using a Malvern® Mastersizer 2000 laser microgranulometer. Organic matter content corresponds to ash free dry weight. Weight-loss after combustion (450 °C, 5h) of lyophilised samples is measured.

3. Results

3.1. Sediment characteristics

The large difference in the continental shelf features between the eastern and western French Mediterranean coast has an important impact on the sediment characteristics observed at our sampling stations.

The 31 stations sampled have been chosen according to the Water Framework Directive strategy, and are systematically positioned within 1 mile from the coast line. Because of this sampling policy and the heterogeneity of the French Mediterranean coast, there is a clear difference in the average water depth of the stations from the western part of our study area (18 m on average) compared to those from the east (40 m on average). The limit between the two areas is approximately positioned between the stations Fos and Carry (Fig. 2a, Appendix B).

The grain size analyses show a clear difference between western and eastern stations (Fig. 2c–e, Appendix B). Stations west of Carry contain a low proportion of sand >250 μm, with the exception of the stations Collioure and Cerbère, which are located at the most western part of the French coast. Conversely, the eastern stations show a high proportion of medium (250–500 μm) and coarse (500–1000 μm) sands, with the exception of some stations (e.g. Marseille Jetée, Ile Embiez, Nice, Menton). There is a significant positive correlation between the percentage of the >500 μm fraction and water depth ($r = 0.50$, $p < 0.05$; Appendix C), which underlines the difference in sediment characteristics along the French Mediterranean coast. Conversely, stations located close to the Rhône river mouth (Fos, Carreau, Beauduc) show a high

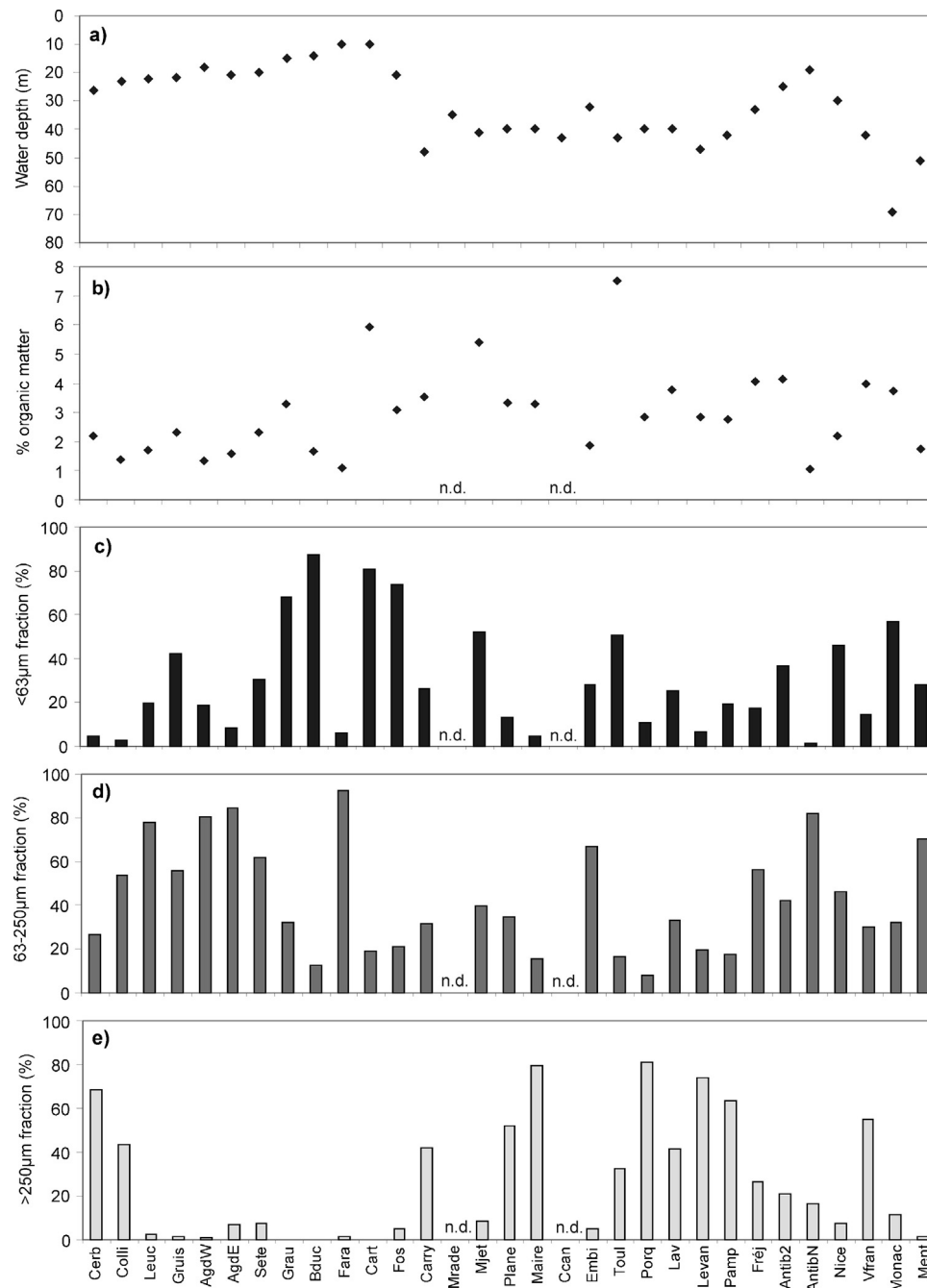


Fig. 2. Environmental parameters for the 31 sampling stations, from west (left side) to east (right side) stations: (a) water depth, (b) percentage of organic matter, (c) percentage of clay and silt (<63 µm), (d) percentage of very fine to fine sand (63–250 µm), and (e) percentage of medium to coarse sand (>250 µm).

proportion of clay and silt particles (<63 µm), in response to a continuous input of fine-grained sediment from the Rhône river.

The organic matter content (Fig. 2b, Appendix B) has been analysed on the total sediment, without any pre-treatment. Consequently, this organic matter is not only composed of marine phytoplankton detritus and of river-supplied continental organic matter, but also by much larger debris of macro-algae and seagrass (roots, leaves). The feeding strategies of foraminifera are various, from detritivory on labile or also more refractory organic matter, to carnivory and bacterivory (review in Murray, 2006). In our study area, marine and continental sedimentary organic matter can probably serve as food for the benthic foraminifera, which is probably not the case for the seagrass debris. In fact, the trophic state of marine sediments is not only dependent on the absolute

quantities of organic matter deposited on the sea floor, but it is also a function of its biochemical composition and nutritional quality for consumers (Pusceddu et al., 2009). Several studies (e.g. Mateo et al., 2006; Østergaard Pedersen et al., 2011) have shown that the roots, rhizomes, and leaf sheaths of *Posidonia* decompose very slowly due to their high content of lignin, cellulose, and phenolic compounds (Harrison, 1989; Klapet et al., 2000), which are not readily degraded by microbes (Godshalk and Wetzel, 1978). Therefore the large amounts of *Posidonia* leaves and roots found at several stations in our study area, resulting in very high OM values in some stations, cannot be considered as readily available food for benthic organisms. Consequently, it appears impossible to use the OM percentages as measure of the trophic level or as an indicator of anthropogenic pressure. There is no clear west-east trend in the

OM percentage (Fig. 2b), but there is a statistically significant positive correlation between the OM content and the percentage of clay/silt ($<63 \mu\text{m}$) particles ($r=0.42$, $p<0.05$; Appendix C), as was observed previously in other coastal areas (e.g. Jorissen, 1987, 1988; Fontanier et al., 2008). Large quantities of macro-algae and sea-grasses (e.g. detritus of *Posidonia* roots) observed in the sediment collected at stations east of Fos explain the abnormally high OM percentages found in some stations with coarse sediments (e.g. Ile Maire, Porquerolles, Ile Levant).

Summarising, natural environmental characteristics appear to be very different between the western stations (lower water depth, fine sediment, enriched in sedimentary organic matter) and the eastern stations (higher water depth, coarser sediment, and sometimes abundant plant remains) in our study area. The faunal assemblages that are colonising these different types of environments will therefore be very different naturally. This bias will have to be taken into account when trying to construct a bio-indicator method based on the foraminiferal faunas. However, it is very

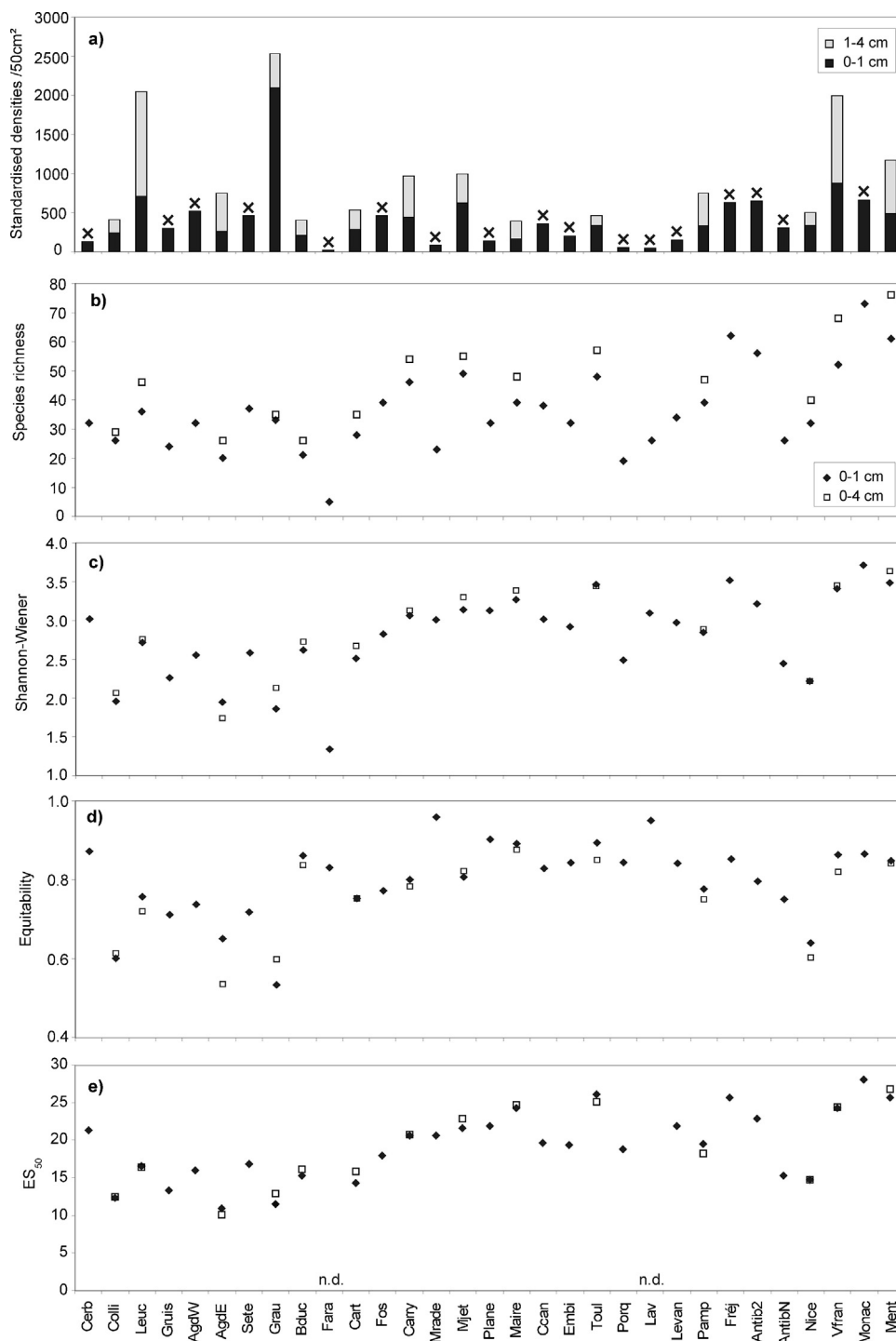


Fig. 3. Density and diversity (species number) of living foraminiferal faunas for the 31 sampling stations, from west (left side) to east (right side), considering either 0–1 cm (black/diamonds) or 0–4 cm (white/squares) sediment intervals: (a) living foraminiferal density (number of specimens standardised for 50 cm², crosses notify samples for which only the first cm of sediment was analysed), (b) species richness, (c) Shannon–Wiener index, (d) equitability index, and (e) ES_{50} . NB: For station Toulon Grande Rade data are based on a study of the 0–1 and 0–3 cm intervals.

probable that this strong west-east dichotomy will equally affect the macrofaunal distribution.

3.2. Diversity and density of the living fauna

Living foraminiferal densities standardised for 50 cm² are highly variable among stations (Fig. 3a, Appendix B). For the 0–1 cm sediment interval (31 stations considered), the total number of foraminifera varies between 22 specimens/50 cm² for station Faraman and 2091 specimens/50 cm² for station Grau du Roi. For the 0–4 cm sediment interval (14 stations considered), total densities vary between 387 and 2526 specimens/50 cm² for stations Ile Maire and Grau du Roi, respectively. The very low densities found at stations Faraman, Lavandou and Porquerolles (22, 43 and 51 specimens/50 cm², respectively) could result from the loss of a large part of the superficial sediment before the Reineck core reached the deck of the ship.

The stations Leucate, Villefranche and Menton exhibit a particularly strong difference in densities between both studied sediment intervals (0–1 and 0–4 cm), indicating the presence of abundant live foraminiferal faunas in deeper sediment layers. In most other stations, this difference is smaller.

Diversity indices are relatively high at all studied stations (Appendix B). Species richness in the first centimetre of sediment varies between 20 (station Agde Est) and 73 species (station Monaco) (Fig. 3b). The Shannon–Wiener index (Fig. 3c) varies between 1.9 (station Grau du Roi) and 3.7 (station Monaco). The equitability index (Fig. 3d), which gives information about the dominance of one or more taxa, varies between 0.53 (station Grau du Roi) and 0.96 (station Marseille Grande Rade). According to these indices, biodiversity seems to increase to the eastern part of the French Mediterranean coast, where the depth of the sampling stations is more important. There is indeed a statistically significant positive correlation between the diversity indices and water depth ($r=0.79$ for ES₅₀, $r=0.74$ for Shannon–Wiener index, $r=0.66$ for specific richness, and $r=0.52$ for equitability index, $p<0.05$ for all correlations; Appendix C).

The expected number of species from a sub-sample of 50 individuals (ES₅₀, Fig. 3e) exhibits smaller differences between stations compared to uncorrected species richness. In general, stations with relatively low total faunal densities (e.g. Cerbère, Marseille Grande Rade or Porquerolles) deviate less from the overall trend. This observation confirms the good performance of the ES₅₀ index in case of samples with large differences in faunal density, which is also the case for the Shannon–Wiener and equitability indices.

The comparison of the diversity indices for the 0–1 cm and 0–4 cm intervals shows first that on average 8 additional species (a maximum of 16 species), have been found when the 1–4 cm interval is added (Fig. 3b). However, the difference in Shannon–Wiener and ES₅₀ indices between the 2 considered depth intervals is relatively small (Fig. 3c and d). Species exclusively found in the 1–4 cm sediment interval are represented by few specimens; the density differences between the 0–1 and 0–4 cm levels highlighted in Fig. 3a are mainly resulting from an increase in the density of species that also occur in the first centimetre of the sediment. The statistical comparison of the diversity indices of both intervals shows a significant difference for the specific richness ($t=-7.05$, $p=0.000$) and Shannon index ($t=-2.71$, $p=0.02$), but no significant differences for the equitability index ($t=1.53$, $p=0.15$) and ES₅₀ ($t=-1.12$, $p=0.28$).

3.3. Vertical distribution of total living foraminiferal faunas

Oxygen profiles have been measured at 11 stations. In fact, overlying water, essential for oxygen profiles, was not always available when we used a Reineck corer. A typical example of an oxygen profile obtained at station Carteau is shown in Appendix D. Oxygen saturation is 93% in the bottom waters and starts to decrease at the sediment–water interface. The oxygen concentration in the interstitial waters decreases rapidly within the first millimetres of the sediment to reach anoxic conditions at 6 mm.

The vertical distribution of living foraminifera is controlled by the oxygen penetration depth in the sediment, the grain size, the availability of labile organic matter and by macrofaunal bioturbation, the latter parameter modifying the former three (e.g. Corliss, 1985; Shirayama, 1984; Corliss and Emerson, 1990).

The vertical distribution of foraminiferal faunas was studied at 14 stations. In order to group these 14 stations in function of sediment grain size, we performed a cluster analysis (using the Ward method) using the different measured grain size fractions (percentages of particles <63 µm, 63–125 µm, 125–250 µm, 250–500 µm and >500 µm). As a result, we obtained two groups of stations: group A with muddy to silty sediments, and group B with sandy sediments. In Table 1, it can be seen that the average living depth (ALD₅/ALD₆) of the live foraminiferal fauna is considerably higher for the sandy stations (group B) than for the clayey-silty stations (group A) (Figs. 4 and 5).

For the stations of group A, with clayey-silty sediment (Fig. 4), faunas present a maximum density in the first centimetre of the sediment (often in the first half centimetre) followed by a

Table 1

Oxygen penetration depth (OPD) and average living depth (ALD₅/ALD₆) for stations where oxygen profiles were performed and average living depth was calculated for cores of 5 or 6 cm length (depending on the slicing) in order to compare the ALD_x of different stations. NB: ALD₃ for Toulon and ALD₄ for Agde Est. Environmental parameters are added to compare with the vertical distribution of foraminifera.

Station	OPD (mm)	ALD ₅ /ALD ₆ (cm)	Depth (m)	%OM	%<63 µm	%63–125 µm	%125–250 µm	%250–500 µm	%>500 µm
Group A									
Grau	7	1.0	15	3.28	67.92	25.86	6.21	0.00	0.00
Toul	–	1.1	43	7.52	50.90	11.59	4.84	2.82	29.85
Cart	6	1.3	10	5.91	80.80	13.07	6.14	0.00	0.00
Mjet	14	1.4	41	5.41	51.91	19.80	19.78	7.61	0.90
Nice	12	1.4	30	2.21	46.12	28.47	17.79	6.80	0.81
Bduc	–	1.6	14	1.68	87.52	10.75	1.73	0.00	0.00
Group B									
Colli	–	1.4	23	1.37	2.89	13.86	39.90	33.74	9.62
AgdE	–	1.8	21	1.57	8.45	27.56	56.99	6.99	0.00
Maire	–	2.0	40	3.31	4.82	4.26	11.22	23.28	56.42
Vfran	–	2.0	42	3.99	14.66	12.25	18.11	22.75	32.22
Leuc	–	2.2	22	1.72	19.73	53.10	24.61	2.41	0.14
Carry	–	2.3	48	3.54	26.28	14.25	17.57	17.72	24.18
Ment	–	2.3	51	1.73	28.26	37.50	32.75	1.49	0.00
Pamp	–	2.7	42	2.78	19.10	6.06	11.39	23.95	39.49

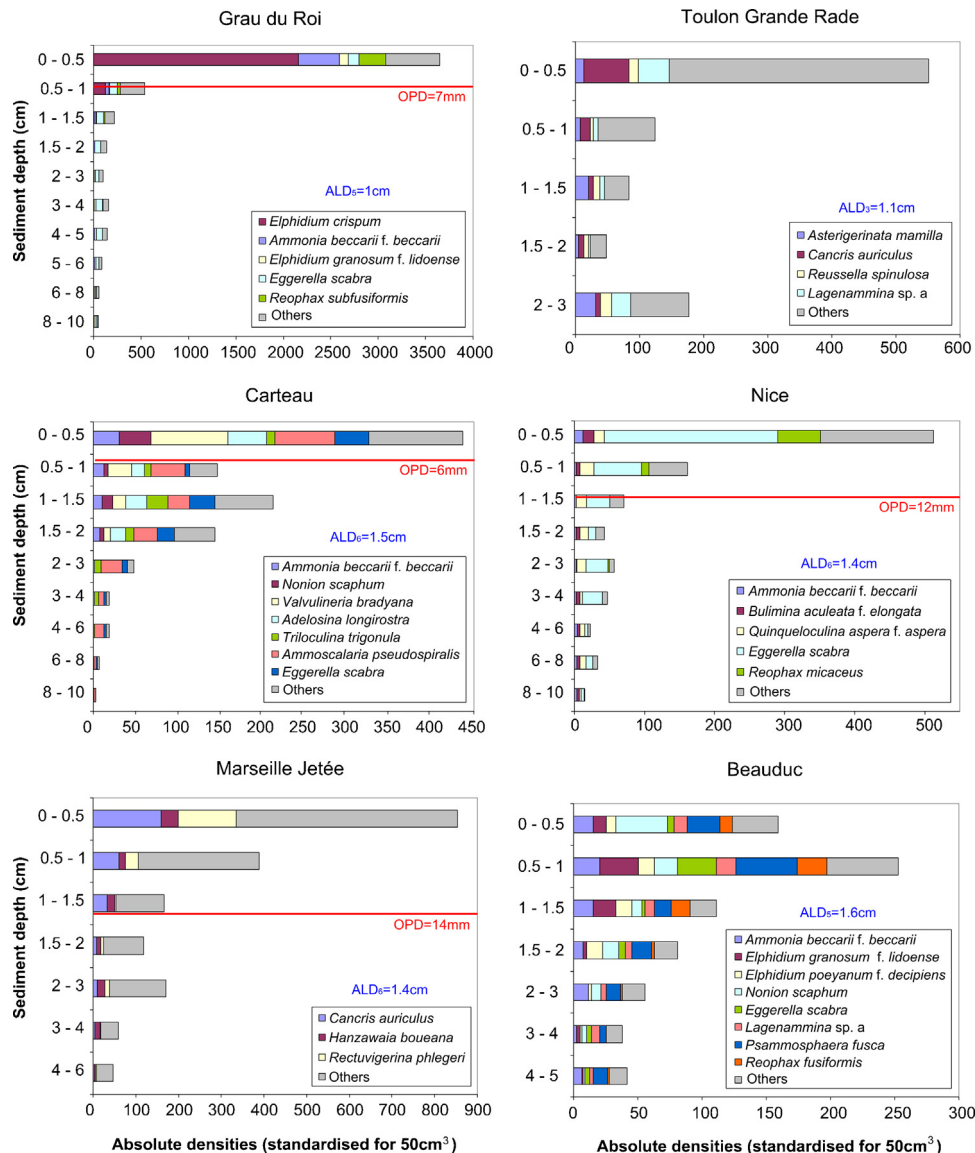


Fig. 4. Vertical distribution of living foraminiferal faunas for stations of group A. Foraminiferal densities are standardised for 50 cm³. Major species (>5% of the total fauna of the core) are presented separately from the rest of the species gathered in “others”. NB: x and y axes scales change according to the stations.

noticeable decrease downcore, more or less sharp. Group A stations are characterised by a relatively shallow average living depth (ALD₅/ALD₆), from 1.0 to 1.6 cm (Table 1). These stations have a relatively high OM content, between 1.68 and 7.52% (4.34% on average). There is a strong negative correlation between the <63 μm particle size fraction and the ALD_x of the total fauna ($r = -0.57$, $p < 0.03$). Generally, silty-clayed marine environments are characterised by weak hydrodynamics allowing the deposition of organic matter (Tyson, 1995) and its adsorption on clay particles (Hedges and Keil, 1995). Fine-grained substrates can therefore often be considered as eutrophic to mesotrophic environments.

The strong surface maximum, together with poor faunas in deeper sediment layers found at these stations is typical for eutrophic environments with limited oxygen penetration depth (a maximum OPD of 14 mm for stations where oxygen profiles were performed) (Jorissen et al., 1995).

Also for the stations of group B (Fig. 5), the foraminiferal vertical distribution is generally characterised by a density maximum in the first centimetre of sediment. However, unlike group A, densities remain high in deeper sediment layers. Consequently, the

ALD₅/ALD₆ of these stations is much higher (between 1.4 and 2.7 cm, 2.1 cm on average; Table 1).

For some stations (e.g. Agde Est, Leucate), the faunal density and composition are almost the same in every sediment layer down to 5 cm. The stations of group B are generally characterised by a lower OM, of 2.5% on average (1.37–3.99%).

Unfortunately, no oxygen measurements could be performed for the stations of group B. However, the abundant faunas in deeper sediment layers suggest that oxygen penetration is considerably deeper here than at the stations of group A, where oxygen penetration varies from 6 to 14 mm (Table 1).

3.4. Species composition of living foraminiferal faunas

In total, 40 major species (>5% in at least one station, 150–500 μm) have been identified: 20 perforate, 8 porcelaneous and 12 agglutinated taxa (Table 2, see Plates 1–4 in Supplementary material 1 for plates showing MEB pictures of major species, Supplementary material 2 for standardised counting data and Supplementary material 3 for the taxonomical list of major species).

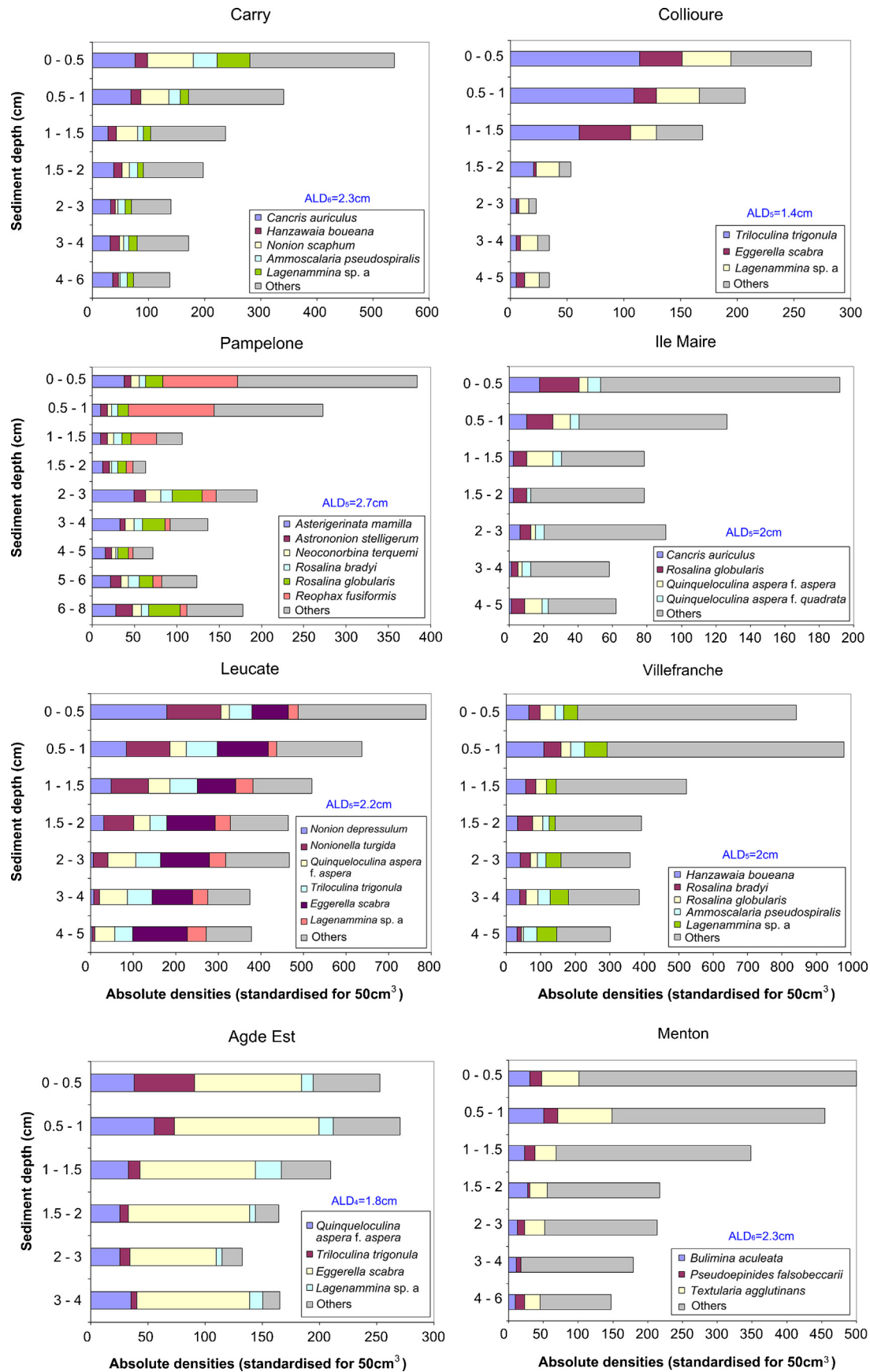


Fig. 5. Vertical distribution of living foraminiferal faunas for stations of group B. Foraminiferal densities are standardised for 50 cm³. Major species (>5% of the total fauna of the core) are presented separately from the rest of the species gathered in "others". NB: x and y axes scales change according to the stations.

Table 2

List of major species (relative density >5% in at least one of the stations studied between 0 and 1 cm).

Perforate species		Porcelaneous species		Agglutinated species	
Name	Abbrev.	Name	Abbrev.	Name	Abbrev.
<i>Ammonia beccarii</i> f. <i>beccarii</i>	Abecc	<i>Adelosina longirostra</i>	Along	<i>Ammoscalaria pseudospiralis</i>	Apseudo
<i>Asterigerinata mamilla</i>	Amami	<i>Biloculinella irregularis</i>	Birreg	<i>Eggerella scabra</i>	Escab
<i>Astronion stelligerum</i>	Astel	<i>Quinqueloculina aspera</i>	Qasp	<i>Lagenammina</i> sp. a	LagenamA
<i>Buccella granulata</i>	Bgran	<i>Quinqueloculina bosciana</i>	Qbosc	<i>Lagenammina</i> sp. b	LagenamB
<i>Bulimina aculeata</i>	Bacul	<i>Quinqueloculina costata</i>	Qcost	<i>Leptohalysis scotti</i>	Rscot
<i>Cancris auriculus</i>	Cauri	<i>Quinqueloculina seminula</i>	Qsemi	<i>Psamosphaera fusca</i>	Pfusc
<i>Cibicides lobatulus</i>	Cloba	<i>Sigmoilina grata</i>	Sgrata	<i>Reophax fusiformis</i>	Rfusif
<i>Elphidium crispum</i>	Ecris	<i>Triloculina trigonula</i>	Ttrigo	<i>Reophax micaceus</i>	Rmica
<i>Elphidium granosum</i>	Egran			<i>Reophax scorpiurus</i>	Rscorp
<i>Elphidium poeyanum</i> f. <i>decipiens</i>	Epoey			<i>Reophax subfusiformis</i>	Rsubfus
<i>Hanzawaia boueana</i>	Hboue			<i>Textularia agglutinans</i>	Taggl
<i>Neonorbina terquemi</i>	Nterq			<i>Textularia sagittula</i>	Tsagit
<i>Nonion depressulum</i>	Ndepres				
<i>Nonion scaphum</i>	Nscap				
<i>Nonionella turgida</i>	Nturg				
<i>Planorbulina mediterraneensis</i>	Pmedit				
<i>Rectuvigerina phlegeri</i>	Rphle				
<i>Rosalina globularis</i>	Rglob				
<i>Spirillina</i> sp.	Spiril				
<i>Valvulinera bradyana</i>	Vbrad				

The relative densities of these major species do not show a statistically significant difference between the 0–1 and 0–4 cm levels (Appendix E). This result indicates that the percentages of the dominant taxa of the first centimetre can be considered as representative for the whole fauna. The following discussion is therefore uniquely based on the 0–1 cm level.

Among the 40 major species, 10 are very common in the study area, and are present in more than 70% of the stations: 2 perforate taxa (*Ammonia beccarii*, *Buccella granulata*), 4 porcelaneous taxa (*Adelosina longirostra*, *Quinqueloculina aspera*, *Quinqueloculina seminula*, *Triloculina trigonula*) and 4 agglutinated taxa (*Eggerella scabra*, *Lagenammina* spp., *Reophax fusiformis*, *Textularia agglutinans*). More specifically, *Eggerella scabra* is clearly the most common species since it is present in 26 of the 31 stations; 15 stations with relative densities over 5% and 7 stations where it represents more than 30% of the total fauna.

On the contrary, some stations are characterised by a strong relative abundance of species that are not frequently found at other stations. For example, *Elphidium crispum* is dominant at the station Grau du Roi where it represents 54.6% (1142 specimens per 50 cm²). This species shows only very low abundances (less than 15 specimens) in 13 other stations and is absent in the rest of the stations. Station Leucate presents also a peculiar faunal composition compared to other studied stations with high relative densities of *Nonion depressulum* (18.4%) and *Nonionella turgida* (16.1%), these species being very scarce in other locations except at Grau du Roi. Leucate is also characterised by a relative abundance of 6.7% of *Leptohalysis scotti*, which appears only with single individuals in 3 other stations.

The analysis of the correlations between the available environmental data and the relative densities of the major species is given in Appendix F. Since our study concerns a very large area with strongly contrasting environmental characteristics, it may be expected that also the faunal composition will show large differences between stations. For example, the positive correlation of *Nonion depressulum* and *Nonionella turgida* with the 63–125 µm grain size fraction is mainly determined by their high percentages at station Leucate, which is characterised by 53% of very fine sand (63–125 µm). At other stations with similar sediment grain size (e.g. Ile Embiez, Agde Ouest), these taxa are much less frequent or absent. Consequently, it becomes difficult to work with individual (marker) species and to observe clear relations between single species percentages and environmental parameters. It is therefore

more relevant to define groups of species with a similar distribution, which will respond in the same way to the environmental parameters.

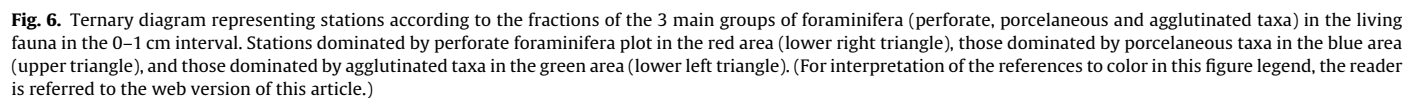
Several trials with Q- and R-mode multivariate statistics (Principal Component Analysis, cluster analysis) to construct species clusters only yielded very inconclusive results. Q-mode PCA results show that *Elphidium crispum* and *Eggerella scabra* are responsible for most of the variability in the dataset when considering the two first PCA axes (see Supplementary material 4). This is due to the strong dominance of *Elphidium crispum* at station Grau du Roi and the high relative densities of *Eggerella scabra* at a number of stations. These species also stand out in the R-mode PCA. The other species cluster together, and do not form clear species groups, even not when considering the next axes. Faunal clusters systematically contain a mix of species with different ecological characteristics, and were therefore very difficult to interpret ecologically (see Supplementary material 5). For this reason, we preferred to test three a priori groupings, based on: (1) wall structure, (2) life position (epiphytic species), and (3) literature observations on tolerance/sensitivity with respect to eutrophication.

3.5. Species groups indicative of environmental quality

According to the comparison between 0–1 cm and 0–4 cm sediment intervals for density, diversity and species composition (see paragraph 5.1 for more details), we considered only data from the first centimetre of the sediment for the study of groups of indicative species of environmental quality.

3.5.1. Species groups according to wall structure

A ternary diagram (Fig. 6, after Murray, 1973) presents the contribution of the 3 main groups (defined by wall structure) to the foraminiferal faunas (of the 0–1 cm level): perforate, porcelaneous and agglutinated species (see also Appendix B). Station Collioure is the only one showing a majority of porcelaneous taxa (Fig. 6, upper blue triangle). The faunas of stations Toulon Grande Rade, Marseille Grande Rade, Ile Plane, Porquerolles, Marseille Jetée, Carry, Fos and Grau du Roi are composed in majority of perforate foraminifera (Fig. 6, lower right red triangle) whereas stations Nice, Agde Est and Ouest, Gruissan, Sète and Ile Embiez are characterised by a dominance of agglutinated tests (between 54 and 71%; Fig. 6, lower left green triangle). At station Beauduc, where porcelaneous taxa are almost absent, equal amounts of perforate and agglutinated taxa



We performed a canonical correspondence analysis to compare the available environmental parameters (grain size fractions, OM content and water depth) with the percentage of the three wall structure groups. The result shows that the five distinguished grain size fractions are distributed in a horse shoe pattern

(Fig. 7). The percentage of porcelainous taxa is plotted in the same area as medium sand (250–500 μm), and is opposed to the percentages of clay and silt. In fact, there is a significant positive correlation between the percentage of porcelainous taxa and the fine and medium sand fractions (for 125–250 μm , $r=0.57$, $p<0.05$; for 250–500 μm , $r=0.55$, $p<0.05$) and a negative correlation with the clay/silt fraction ($r=-0.71$, $p<0.05$; Appendix C).



The percentage of perforate foraminifera plots in the same area as OM content and water depth; there is indeed a positive correlation between their percentage and the percentage of clay/silt ($r=0.42$, $p<0.05$) and with the OM content ($r=0.50$, $p<0.05$; Appendix C). Finally, the percentage of agglutinated taxa plots together with the 63–125 μm fraction showing a positive correlation ($r=0.47$, $p<0.05$). This group anti-correlates with coarse sand ($>500 \mu\text{m}$, $r=-0.52$, $p<0.05$). In general, the distribution of this group seems to be opposite to the one of the perforate taxa ($r=-0.74$, $p<0.05$; Appendix C).

3.5.2. Species group according to life position (epiphytic species)

To constitute the epiphytic species group (i.e. capable to live fixed on algae), we selected the species classified in morphotypes A and B as defined by Langer (1993). These morphotypes have been defined according to the different modes of surface attachment and the feeding strategies. Morphotype A represents stationary, permanently attached species which secrete an organic substance to glue to seagrass leaves or algal blades (e.g. *Planorbulina mediterraneensis*). Morphotype B represents temporary attached species which have a trochospiral shape with apertures facing the substrate (e.g. *Rosalina globularis*). Morphotypes C and D are not considered in our group of epiphytic species since they can also live in areas without seagrass or algae considering their permanently motile behaviour (e.g. elphidiids, porcelaneous species). The epiphytic species

identified in our samples are *Asterigerinata mamilla*, *Cibicides lobatulus*, *Gavelinopsis praegeri*, *Hanzawaia boueana*, *Neonorbina terquemi*, *Planorbulina mediterraneensis*, *Rosalina bradyi*, *Rosalina globularis*, *Rosalina vilardeboana* and other *Rosalina* species (e.g. Jorissen, 1987; Kitazato, 1988; Langer, 1993; Barmawidjaja et al., 1995; Schönfeld, 2002; Murray, 2006; Buosi et al., 2012). These epiphytic species are indicative of the presence of vegetation in the vicinity of the sampling station and generally of a good ventilation of bottom waters. According to van der Zwaan et al. (1999), many epiphytic species are sensitive to oxygen-limited conditions and would be competitive in oligotrophic environments. They are mainly found in sandy sediments (Pujos, 1976; Spindler, 1980; Bizon and Bizon, 1984; Jorissen, 1987; Murray, 1991; Villanueva Guimerans and Cervera Currado, 1999; Mendes et al., 2004; Mojtabah et al., 2006) and some of these species, such as *Cibicides lobatulus* and *G. praegeri*, can tolerate high energy environments (Coppa and Di Tuoro, 1995; Guimerans and Currado, 1999; Schönfeld, 2002; Panieri et al., 2005; Martins et al., 2007; Milker et al., 2009). In our study area, the *Posidonia* meadows provide abundant niches for these foraminiferal species; the rhizomes act as sediment traps and the leaves are often colonised by motile or (temporarily) fixed epiphytic foraminifera (Vénec-Peyré, 1984; Langer, 1993).

We calculated the cumulative percentage of epiphytic species for the 0–1 cm interval of the 31 studied stations (Fig. 8a, Appendix

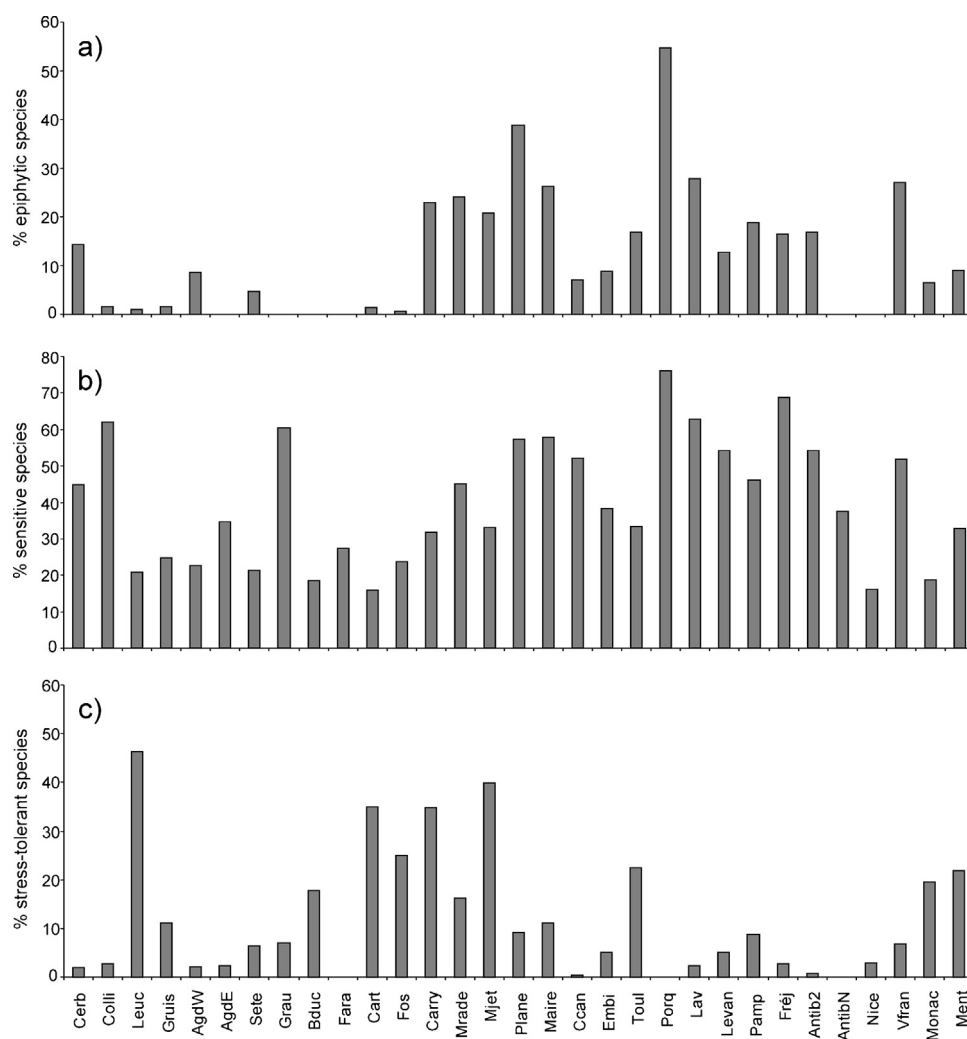


Fig. 8. Percentage of indicative species in the sample (0–1 cm interval) at each station (West-East transect): (a) epiphytic species, (b) sensitive species, and (c) stress-tolerant species.

B). Fig. 8a highlights again the clear difference between western shallow stations and eastern deeper stations (limit between Fos and Carry), with the exception of Antibes Nord and Nice where epiphytic species are absent. As illustrated by the CCA analysis (Fig. 9), there is a positive correlation between the percentage of epiphytic species and water depth ($r=0.53$, $p<0.05$), medium and coarse sediment (for 250–500 μm , $r=0.40$, $p<0.05$; for $>500 \mu\text{m}$, $r=0.80$, $p<0.05$). On the other side, there is a negative correlation with fine sands (63–125 μm ; $r=-0.50$, $p<0.05$; Appendix C).

3.5.3. Species groups according to tolerance/sensitivity to organic enrichment

According to the literature, we defined two species groups: (1) a group of “stress-tolerant” species, with a high percentage being indicative of stressed conditions, such as eutrophication or abundant supplies of fine-grained sediments, and (2) a group of sensitive species, which are supposed to be indicative of a good overall quality of the ecosystem, and which should disappear when environmental conditions become more stressful.

Ten stress-tolerant taxa were identified on the basis of literature evidence: *Bulimina* spp., *Cancris auriculus*, *Nonion scaphum*, *Nonion depressulum*, *Nonionella turgida*, *Nonionella stella*, *Pseudopionides falsobecarii*, *Rectuvigerina phlegeri*, *Valvulinaria bradyana* and the agglutinated *Leptohalysis scotti* (see Plate 2 in Supplementary material 1). The observations presented in the literature which supported our decision to place these 10 taxa in the tolerant group are listed in Appendix G.

The group of sensitive species (see Plate 3 and 4 in Supplementary material 1) includes all porcelaneous species and all epiphytic species. In addition, we also included other motile epiphytic species (morphotypes C and D according to Langer, 1993) such as *Elphidium* species (*E. crispum*, *E. granosum* and *E. poeyanum*), *Reussella spinulosa* and *Spirillina* spp. According to the literature that supports our choice to group all these species sensitive to stressed conditions (see Appendix G for literature references on which this grouping was based), a poor representation of this group in the total fauna

would be indicative of enrichment in muddy sediments, eventually leading to low oxygen conditions.

Among the 40 major species identified, 14 species have not been assigned to one of these two groups, either because they are neither sensitive nor stress-tolerant, or due to a lack of well documented studies with clear pollution gradients, or due to contradictory literature data with respect to their ecological characteristics.

The case of *Eggerella scabra* is particularly striking. Although this species has been reported in several articles as being able to tolerate stressed conditions, we did not include it in the group of tolerant species. *Eggerella scabra* is a continental shelf species (e.g. Murray, 1991; Barmawidjaja et al., 1992) that lives in muddy to sandy substrates (Murray, 1986; Alve and Nagy, 1986; Scott et al., 2003) and in various microhabitats, from the oxygenated sediment surface to the deepest anoxic layers (e.g. Barmawidjaja et al., 1992; Jorissen et al., 1992; Ernst et al., 2002, 2005; Duijnste et al., 2003, 2004). It appears therefore to be tolerant for hypoxic conditions. For instance, *Eggerella scabra* is common in the Adriatic Sea, in areas where important amounts of degraded organic matter cause oxygen depletion (Donnici and Serandrei-Barbero, 2002). It has also been shown to support extremely polluted environments in Sjøfjord, western Norway (Alve, 1991). On the other hand, this species is very common and typical in many apparently unpolluted coastal Mediterranean environments (e.g. Vénec-Peyré, 1984; Donnici and Serandrei-Barbero, 2002; Duijnste et al., 2003; Frontalini and Coccioni, 2008; Mojtahid et al., 2009; Goineau et al., 2012; Sabbatini et al., 2010, 2012). Several authors suggested that this species has a poorer tolerance to stressed conditions than some clear opportunists, although it has a great ability to withstand fluctuations in diverse parameters including an absence of labile organic matter (Scott et al., 2003; Mojtahid, 2007; de Nooijer et al., 2008; Sabbatini et al., 2012). Finally, some authors have considered *Eggerella scabra* as an epiphytic species on seagrass (Redois and Debenay, 1996; Debenay, 2000), again suggesting that it can be a dominant faunal element in high quality ecosystems. Because of this strongly contrasting evidence and the high densities of *Eggerella scabra* in most

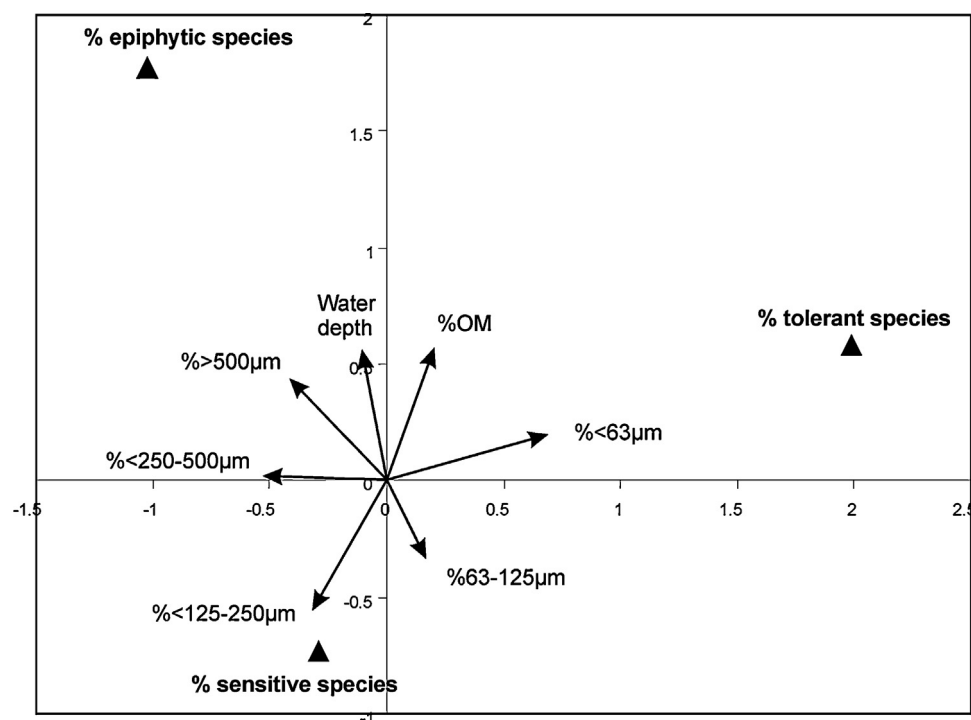


Fig. 9. Canonical correspondence analysis (axis 1 vs. axis 2) performed on environmental parameters and the percentages of the indicative species groups: epiphytic species, stress-tolerant and sensitive species (without considering Marseille Grande Rade and Cap Canaille for which no environmental data were available).

of our studied stations, we decided not to include this species in our stress-tolerant group so that it does not obscure the message given by more clear stress-tolerant species.

Fig. 8b and c shows the percentages of sensitive and stress-tolerant species in our study area following a West-East transect. In our dataset, the percentage of stress-tolerant species positively correlates with the percentage of fine particles ($r=0.48$, $p<0.05$) and organic matter ($r=0.40$, $p<0.05$; Fig. 9, Appendix C). Conversely, stress-tolerant species are weakly represented in eastern stations, in spite of high organic matter contents measured at some stations (e.g. stations Fréjus and Antibes).

Conversely, sensitive species are negatively correlated with the percentage of fine particles (for $<63\ \mu\text{m}$, $r=-0.49$, $p<0.05$; for $63\text{--}125\ \mu\text{m}$, $r=-0.44$, $p<0.05$) and positively correlated with coarser particles (for $250\text{--}500\ \mu\text{m}$, $r=0.59$, $p<0.05$; for $>500\ \mu\text{m}$, $r=0.60$, $p<0.05$; Fig. 9, Appendix C).

4. Discussion

4.1. Representativity of the fauna of the first centimetre of sediment

Ecological studies of recent foraminiferal faunas are usually based on the analyses of the total fauna in the sediment column, down to 5 or 10 cm depth. In fact, the vertical distribution of foraminifera can give information about the ecological strategies of different species or about the environmental conditions. In our study, the fauna of deeper sediment layers allows us to distinguish two types of environments. More eutrophic, silty/clayey stations with a limited oxygen penetration depth have the large majority of the fauna in the topmost centimetre, whereas sandy stations, probably with lower organic matter supplies, show a more even faunal distribution in the first 2 to 5 cm of the sediment.

Although this environmental information is not without interest, the significantly longer picking time required to obtain data from deeper layer makes that the study of the vertical distribution analysis is hardly possible for bio-monitoring studies, in which economical aspects are important, and strict deadlines have often to be respected. Recently, the FOBIMO group recommended therefore to limit foraminiferal bio-monitoring studies to the analysis of the first centimetre of the sediment. This recommendation was supported by the results of Bouchet et al. (2012), who studied the faunal response to various oxygen concentrations in the Norwegian Skagerrak, using diversity indices based on the faunas in the 0–1 cm and 0–2 cm intervals. It turned out that the results were virtually similar, suggesting that the study of the 0–1 cm was sufficient.

However, before taking the decision to restrict the faunal analysis to the uppermost centimetre, we wanted to verify whether this does not lead to an erroneous or incomplete interpretation of the faunal response to environmental conditions when considering our coastal Mediterranean samples. In our study area, species living exclusively in deeper sediment layers (e.g. Corliss, 1985; Jorissen et al., 1995) were not observed (Figs. 4 and 5). Buzas et al. (1993) highlighted the fact that the microhabitat succession usually observed in deep water (outer continental shelf and slope) is much less evident on inner continental shelf environments. They explained this difference by the more dynamic nature of coastal areas (sediment disturbance, bioturbation, etc.).

To know if a study restricted to the first centimetre of sediment (generally containing the majority of the living fauna) is sufficient to correctly define the environmental quality, we compared faunal parameters between 0–1 and 0–4 cm intervals. Statistical comparison of equitability indices and ES_{50} show no significant difference between 0–1 and 0–4 cm. However, faunal densities are significantly different. For the Shannon index, the statistical test identified

significantly higher values for the 0–4 cm interval (test based on the sign of the difference). However, the differences are small (average shift between the values from 0–1 and 0–4 cm intervals of 0.11), and would not cause major changes in the classification of the stations into the different quality categories. It is also interesting to observe that there are no significant differences in the relative densities of major species which change only slightly between the 0–1 and 0–4 cm intervals (Wilcoxon test, Appendix E).

In view of all these results, we conclude that in our study area, the first centimetre of the sediment gives a very good picture of the overall live fauna, its diversity and composition. Therefore, our results fully support the recommendation of the FOBIMO group (Schönfeld et al., 2012).

4.2. Relevance of the various faunal parameters for ecosystem quality evaluation

Ideally, the development of a faunal index of environmental quality should be based on a precise knowledge of pollution sources and intensities in the study area. The analysis of faunal patterns along a well-described pollution gradient makes it possible to distinguish species with various degrees of tolerance, and to identify the faunal parameter(s) or indices that correlate best with the state of the environment, as defined by the concentration of one or more pollutants. Usually, such studies focus on the impact of a single stress parameter on the foraminiferal faunas, such as bottom water oxygen concentration (Bouchet et al., 2012), eutrophication (Mojtahid et al., 2008), or chemical pollution (Frontalini and Coccioni, 2008; Mojtahid et al., 2006). In our study, the geographical area of concern is very wide, pollutants are dispersed in an erratic way, and their concentration is not known. Consequently, we do not dispose of a clear transect following a pollution gradient, and our approach has therefore to be slightly different. We cannot have the ambition to directly develop a biotic index, but instead, we will try to determine which faunal parameters could be relevant to adequately describe the ecosystem health.

4.2.1. Biodiversity indices

According to different diversity indices calculated, biodiversity seems to be higher at the eastern part of the French Mediterranean coast. Unfortunately, because of the strong positive correlation between diversity indices and water depth, it is impossible to say whether the higher values of the diversity indices of the eastern stations indicate a higher overall biodiversity, or whether they are the result of a sampling bias (shift in water depth).

Diversity indices give important information about biodiversity and faunal equilibrium at a station. It has been shown that biodiversity indices may be useful to classify the ecosystem quality in strongly polluted conditions (e.g. Bouchet et al., 2012; Armynot du Chatelet et al., 2004). Our study area differs from the severely stressed environments described by these authors, because of the absence of a clear stress parameter, such as oxygen depletion or heavy metal pollution. In fact, the French Mediterranean coast is generally considered as rather oligotrophic (e.g. Bosc et al., 2004). Consequently, a slight eutrophication of the benthic ecosystem does not necessarily lead to a decreased biodiversity, but could easily cause an increase of the values of diversity indices. Based on several studies of macrofauna along a gradient of organic enrichment, the Pearson–Rosenberg model (also called SAB model, Pearson and Rosenberg, 1978) clearly shows that a slight increase in organic matter content leads first to an increase of the number of species. Several earlier studies show that foraminiferal diversity is decreasing along bathymetric transects in response to lowering of the OM flux towards greater water depth (e.g. Rathburn et al., 1996; Schmiedl et al., 2000; Fontanier et al., 2008). Consequently, we think that in the oligotrophic Mediterranean sea, diversity indices

are not an appropriate tool to describe the environmental quality of benthic ecosystems.

4.2.2. Species groups according to wall structure

According to our data (Fig. 7, Appendix C) and the literature, porcelaneous taxa have the clearest ecological response to environmental change. They are found abundantly in coarse-grained shallow water environments with a low OM content and oxygen-saturated bottom waters (e.g. Jorissen, 1988; Donnici and Serandrei-Barbero, 2002). Bizon and Bizon (1984) observed that this group is also abundant in sandy sediments on the continental shelf off the Rhône River. In our study, their percentage shows a clear decrease with an increasing percentage of fine sediment (<63 µm; Fig. 7, Appendix C). Therefore, samples with a high percentage of porcelaneous taxa should denote stations with rather good environmental quality, whereas the opposite should be true for samples with very low percentages of porcelaneous taxa. For example, stations Collioure, Fréjus and Cap Canaille all show more than 40% of porcelaneous taxa, suggesting healthy environmental conditions. Conversely, stations Grau du Roi, Fos, Carry and Marseille Jetée show very low percentages of porcelaneous specimens (less than 8%), and very high percentages of perforate foraminifera (over 60%) (Fig. 6), suggesting that these stations with clayey-silty sediments may be characterised by a slightly degraded ecological state. However, we cannot push the interpretation much further. In fact, wall structure groups present the disadvantage to separate species according to morphological criteria, which do not necessarily correspond exactly to ecological preferences and tolerances (Buzas et al., 1993). When we look in more detail at the species composing the three groups, it appears that some important species do not at all respect the general tendency. For example, some porcelaneous species have been observed to behave as opportunistic species in particular conditions. So has *Quinqueloculina seminula* been described as an early foraminiferal recoloniser of the benthic ecosystem after a gravity flow in the Whittard canyon (Duros et al., 2011) and on an ash layer deposit around Mt. Pinatubo in the South China Sea (Hess and Kuhnt, 1996). Another example is the group of perforate species which includes species that we classified as tolerant to stressed conditions (e.g. *Nonion scaphum*, *Cancris auriculus*) and epiphytic species which are generally considered as very sensitive to eutrophication.

Summarising, an index based on the cumulative percentages of the three wall structure groups can give a rapid first overall characterisation of the state of the environment, but may in some specific cases lead to erroneous conclusions. It appears therefore that it is more judicious to base a biotic index on groups of indicator taxa which have a similar response to stressed conditions.

4.2.3. Species groups according to life position (epiphytic species)

Our data seem to confirm the literature: epiphytic species are most successful on coarse-grained substrates (Fig. 9, Appendix C), where bottom waters are normally well oxygenated. The rather surprising positive correlation with OM content (Appendix C) is probably caused by the presence of abundant larger plant debris in seagrass meadows, leading to anomalously high OM values. High percentages of epiphytes are found in eastern part of the French Mediterranean coast as well as in front of Banyuls-sur-Mer, where *Posidonia* meadows are growing (Gobert et al., 2009). Since *Posidonia* meadows are known to be highly sensitive to human disturbance (Boudouresque et al., 2000, 2006), our observations suggest that high percentages of epiphytic species could indeed be indicative of a good ecosystem quality. However, their absence at stations naturally characterised by more fine-grained substrates and lack of vegetation cover cannot be interpreted as indicative of a bad ecosystem state. It appears therefore that this parameter can emphasise a very good ecosystem state in some cases, but cannot

be used to characterise the environmental quality along the entire French Mediterranean coast.

4.2.4. Species groups according to tolerance/sensitivity to organic enrichment

Fig. 8c shows a clear increase in the percentage of stress-tolerant species in the stations located around the Rhône River mouth. The Rhône River is the main sediment source in the Gulf of Lions (80%; Durrieu de Madron et al., 2000). Hence, stations located in the vicinity of the delta are influenced by supplies of fine sediment and terrestrial organic matter. However, these stress-tolerant species do not occur in all stations exhibiting high percentages of OM because in several eastern stations these high values reflect the presence of macro-algae detritus, as mentioned earlier.

Sensitive species are less adapted to inhabit muddy to silty substrates which are often characterised by a varying degree of organic enrichment. This enrichment can be entirely natural, or partly, in some cases even entirely anthropogenic. Therefore, a historical disappearance of sensitive species (shown by a comparison of recent and fossil faunas) can highlight either a (natural) shift from more sandy to more muddy sediment, or an increase of anthropogenic organic supplies.

Summarising, the proportions of stress-tolerant and sensitive species can allow us to distinguish two kinds of environments:

- (1) Faunas characterised by a high percentage of stress-tolerant species and a low percentage of sensitive species are indicative of fine-grained substrates often associated with high organic matter contents. This concerns stations Leucate, Gruissan, Beauduc, Carteau, Fos, Carry, Marseille Grande Rade, Marseille Jetée, Toulon Grande Rade, Monaco and Menton. Often, the predominance of stress-tolerant taxa is probably the result of natural conditions. However, in some cases it may be due to a superimposed anthropogenic impact.
- (2) Faunas with a high percentage of sensitive species and a low percentage of stress-tolerant species are indicative of sandy substrates with relatively low organic matter content and well oxygenated bottom waters. Such a situation was encountered at stations Cerbère, Collioure, Agde Ouest, Agde Est, Sète, Grau du Roi, Faraman, Cap Canaille, Ile Plane, Ile Maire, Embiez, Porquerolles, Lavandou, Ile du Levant, Pampelone, Fréjus, Antibes 2, Antibes Nord, Nice and Villefranche.

It appears that the information given by the group of stress-tolerant species is very similar (but opposed) to the information given by the group of sensitive species. However, as shown in Fig. 8b, sensitive species are well represented in all stations (from 16 to 76%) whereas the percentage of stress-tolerant species appears to be more discriminative (from 0 to 46%). This difference is essential for the development of a biotic index of ecological quality status.

Although the group of stress-tolerant species apparently can inform us about the degree of stress at a particular station, it does not tell us whether this stress results entirely from natural conditions or is partly, or totally, due to an anthropogenic impact. Since the aim of bio-monitoring studies is to evaluate the anthropogenic impact on the ecosystem (excluding natural eutrophication); it is absolutely essential to deconvolve these two parameters.

4.3. Correction for natural eutrophication phenomena

In this study, the analyses of the environmental parameters (water depth, organic matter, grain size fraction) highlighted the clear (natural) environmental differences between: (1) stations located on West side of the Rhône River with relatively shallow water depths (18 m depth in average) and clayey to fine sandy sediments; (2) stations located in front of the Rhône river mouth

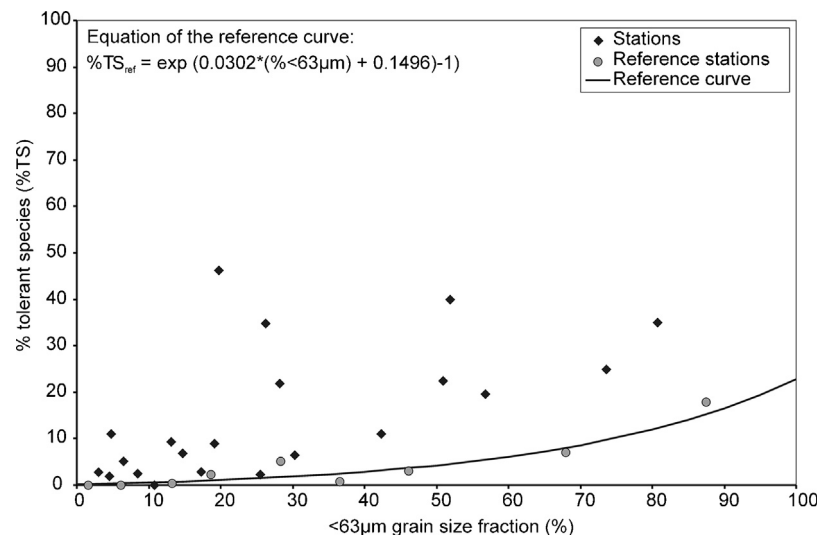


Fig. 10. Percentage of stress-tolerant species versus the percentage of particles <63 μm in the different stations studied along the Mediterranean coast. The exponential curve is thought to represent the percentage of stress-tolerant species in natural conditions (without anthropogenic influence).

and in the Gulf of Fos with clayey sediments enriched in organic matter; and (3) the eastern stations with relatively important water depth (40 m depth in average) and coarser sediments. Living foraminiferal faunas of the 31 analysed stations respond clearly to this natural variability of environmental parameters with changes in species composition. Foraminifera (and benthic fauna in general) are largely influenced by sediment grain size. A simple faunal analysis shows large differences between faunas from clayey and sandy substrates. In fact, faunas living on clayed substrates are more adapted to naturally enriched conditions (eutrophication), often characterised by increased OM concentrations and sometimes

seasonal low oxygen concentrations. For this reason stations with muddy substrates tend to show an elevated proportion of stress-tolerant species, even if the concerned ecosystem is exempt from anthropogenic impact.

We think therefore, that it is necessary to define reference conditions in function of grain size distribution, in order to avoid a basic and erroneous interpretation of faunal data that would consider any station with a clayey substrate of bad quality.

In our database, we selected 8 stations (Agde Ouest, Grau du Roi, Beauduc, Cap Canaille, Embiez, Antibes 2, Antibes Nord and Nice) with different proportions of fine grain-sized sediment (<63 μm

Table 3
Values of the standardised percentage of stress-tolerant species. The parameters required for the calculation of the %TS_{std} are indicated. Data are missing for Marseille Grande Rade due to the lack of particle size measurements at this station.

Stations	%TS _x (%tolerant species, station x)	%<63 μm (<63 μm particles)	%TS _{ref} (%tolerant species, theoretical reference conditions)	%TS _{std} (standardised %tolerant species)
Cerb	1.90	4.56	0.3	1.6
Colli	2.67	2.89	0.3	2.4
Leuc	46.28	19.73	1.1	45.7
Gruis	11.11	42.36	3.2	8.2
AgdW	2.21	18.70	1.0	1.2
AgdE	2.42	8.45	0.5	1.9
Sete	6.54	30.35	1.9	4.7
Grau	7.07	67.92	8.0	−1.1
Bduc	17.79	87.52	15.3	2.9
Fara	0.00	6.16	0.4	−0.4
Cart	35.06	80.80	12.3	25.9
Fos	24.93	73.57	9.7	16.9
Carry	34.77	26.28	1.6	33.7
Mrade	16.13			
Mjet	39.84	51.91	4.6	37.0
Plane	9.26	13.18	0.7	8.6
Maire	11.11	4.82	0.3	10.8
Ccan	0.36	13.25	0.7	−0.4
Embi	5.10	28.30	1.7	3.4
Toul	22.47	50.90	4.4	18.9
Porq	0.00	10.83	0.6	−0.6
Lav	2.33	25.56	1.5	0.8
Levan	5.08	6.47	0.4	4.7
Pamp	8.85	19.10	1.1	7.9
Fréj	2.82	17.35	1.0	1.9
Antib2	0.78	36.56	2.5	−1.8
AntibN	0.00	1.62	0.2	−0.2
Nice	3.00	46.12	3.7	−0.7
Vfran	6.93	14.66	0.8	6.2
Monac	19.51	56.83	5.5	14.9
Ment	21.96	28.26	1.7	20.6

size fraction), which show minimal percentages of stress-tolerant species. These stations were used to define the reference faunas, in other words, the percentages of stress-tolerant species expected to be found in a natural environment with a certain grain-size composition, without any anthropogenic impact (Fig. 10). If only very few (1 or 3) reference stations are selected, there is the risk that they do not represent correctly all the environmental conditions of the study area. Our method, based on 8 reference stations, represents a wide range of coarse sand to clayey substrates, with many intermediate conditions being represented. The stations Faraman, Lavandou and Porquerolles, which also present very low percentages of tolerant species were not retained as reference stations because of their low total number of individuals (<51 ind.) which make them statistically less robust.

Knowing the theoretical percentage of tolerant species in reference conditions for each grain size (defined by the equation $\%TS_{ref} = \exp(0.0302 \times (\%<63 \mu\text{m}) + 0.1496) - 1$), it is then possible to calculate the standardised percentage of tolerant species ($\%TS_{std}$) using the following formula, for a given grain size composition:

$$\%TS_{std} = \frac{\%TS_x - \%TS_{ref}}{100 - \%TS_{ref}} \times 100$$

where $\%TS_x$ is the percentage of tolerant species at station x , and $\%TS_{ref}$ is the theoretical percentage of tolerant species expected at a station with a certain proportion of <63 μm particles, in the absence of anthropogenic impact (c.f. exponential curve equation).

The $\%TS_{std}$, which varies from 0 to 100, describes the increase of the number of stress-tolerant taxa with respect to a reference station with a similar grain-size. Exceptionally, some stations can present a lower percentage of stress-tolerant species than the reference conditions, leading to negative values (Table 3, Fig. 10). Values of $\%TS_{std}$ close to 0 are indicative of a very high environmental quality, whereas values close to 100 would indicate a very high anthropogenic impact.

The standardised percentages of tolerant species for 30 studied stations are presented in Table 3 (except for station Marseille Grande Rade for which we do not have a grain size analyses). Twenty-one stations out of 30 show a $\%TS_{std}$ below 10%, suggesting that the ecological quality at these stations is high, close to theoretical reference conditions. The other 9 stations contain a $\%TS_{std}$ between 10 and 50, indicating that the percentage of stress-tolerant species is higher than would be expected in natural conditions. This concerns particularly the stations Carry, Marseille Jetée and Leucate stations which exhibit a $\%TS_{std}$ higher than 30%. The benthic foraminiferal faunas of these stations are very probably impacted by human activities.

5. Conclusion

In the literature, the study of foraminiferal faunas along the French Mediterranean coast is rather dispersed, with some older

studies dealing with total (dead and living individuals) assemblages (Blanc-Vernet, 1969; Bizon and Bizon, 1984; Vénec-Peyré, 1984), and some more recent studies on living foraminifera around the Rhône river mouth (Mojtahid et al., 2009, 2010; Goineau et al., 2011, 2012). Our study of 31 stations presents for the first time a description of living (Rose Bengal stained) foraminiferal faunas along the entire French Mediterranean coast except Corsica.

The comparative study, for 14 stations, of two different sediment intervals, 0–1 cm and 0–4 cm, clearly shows that the analysis of the uppermost centimetre of sediment is sufficient to obtain relevant information needed for bio-monitoring purposes. In our sandy to silty coastal area, intermediate to deep infaunal species are virtually absent so that the faunal composition of the topmost centimetre is representative of the whole sediment column. This conclusion strongly supports the recommendation of the FOBIMO group (Schönfeld et al., 2012).

Our analysis of the different faunal parameters led us to the conclusion that the use of indicator species, such as stress-tolerant or sensitive species, is more relevant than the use of diversity indices for the evaluation of ecosystem quality, at least in rather oligotrophic areas such as the Mediterranean Sea. Finally, we propose a method to distinguish between natural and anthropogenic eutrophication phenomena by determining the expected percentage of stress-tolerant taxa in natural environments, in function of sediment grain size, and by correcting the observed percentage of stress-tolerant taxa accordingly. This study is a first step towards the development of a foraminiferal index of ecosystem quality for the coastal Mediterranean Sea that could be used in the context of the Marine Strategy Framework Directive (2008/56/EC). Our index has to be tested at other stations, ideally located on a gradient of disturbance. Furthermore, some aspects deserve to be further explored, such as the pertinence of our list of tolerant species in other Mediterranean coastal areas, the potential of the comparison of live and dead faunas to select indicator species, or the relevance of a multimetric index (cf., M-AMBI) combining indicator species and diversity indices.

Acknowledgements

We are grateful to the crew of *Europe* vessel (IFREMER) and scientists who participated to the oceanographic cruise. We also thank Mélissa Gauthier (Angers University) who helped for laboratory sample treatments. Finally, we would like to thank the Agence de l'Eau Rhône Méditerranée et Corse for funding this study.

Appendix A.

Localisation (WGS84) and water depth of the stations. The sediment layers analysed for living foraminiferal faunas are indicated.

Station	Station abbr.	Longitude (°E)	Latitude (°N)
Cerbère	Cerb	3°10'21"	42°26'43"
Collioure	Colli	3°05'22"	42°31'54"
Leucate	Leuc	3°04'00"	42°51'09"
Gruissan	Gruis	3°12'16"	43°09'12"
Agde Ouest	AgdW	3°28'16"	43°14'21"
Agde Est	AgdE	3°32'23"	43°16'17"
Sète	Sete	3°42'41"	43°22'38"
Grau du Roi	Grau	4°03'12"	43°31'34"
Beauduc	Bduc	4°30'08"	43°24'84"
Faraman	Fara	4°43'13"	43°20'00"
Carteau	Cart	4°53'44"	43°23'08"
Fos	Fos	4°55'46"	43°21'38"
Carry	Carry	5°09'38"	43°18'40"
Marseille Grande Rade	Mrade	5°18'28"	43°16'10"
Marseille Jetée	Mjet	5°19'41"	43°20'15"
Marseille-Ile Plane	Plane	5°23'02"	43°11'41"
Ile Maire	Maire	5°20'50"	43°12'16"
Cap Canaille	Ccan	5°33'11"	43°11'07"
Ile Embiez	Embi	5°46'47"	43°06'08"
Toulon Gde Rade	Toul	5°57'54"	43°05'34"
Porquerolles	Porq	6°16'28"	43°01'08"
Lavandou	Lav	6°23'13"	43°06'08"
Ile Levant	Levan	6°25'60"	43°00'13"
Pampelone	Pamp	6°41'44"	43°13'44"
Fréjus	Fréj	6°52'07"	43°25'20"
Antibes 2	Antib2	7°08'29"	43°33'34"
Antibes Nord	AntibN	7°08'07"	43°36'47"
Nice Ville	Nice	7°14'08"	43°40'51"
Villefranche	Vfran	7°18'40"	43°41'35"
Monaco 2	Monac	7°25'47"	43°43'43"
Menton	Ment	7°59'41"	43°45'21"

Station	Water depth (m)	Sediment interval studied (cm), living fauna	Oxygen profiles
Cerbère	26	0–1	
Collioure	23	0–5	
Leucate	22	0–5	
Gruissan	21.5	0–1	
Agde Ouest	18	0–1	
Agde Est	21	0–4	
Sète	20	0–1	
Grau du Roi	15	0–10	X
Beauduc	14	0–5	
Faraman	10	0–1	
Carteau	10	0–10	X
Fos	20.8	0–1	X
Carry	48	0–6	
Marseille Grande Rade	35	0–1	
Marseille Jetée	41	0–6	X
Marseille-Ile Plane	40	0–1	
Ile Maire	40	0–5	
Cap Canaille	43	0–1	
Ile Embiez	32	0–1	
Toulon Gde Rade	43	0–3	
Porquerolles	40	0–1	
Lavandou	40	0–1	
Ile Levant	47	0–1	
Pampelone	42	0–8	
Fréjus	33	0–1	X
Antibes 2	25	0–1	X
Antibes Nord	19	0–1	
Nice Ville	30	0–10	X
Villefranche	42	0–5	
Monaco 2	69	0–1	X
Menton	51	0–6	

Appendix B.

Environmental parameters and faunal parameters (considering foraminiferal faunas from the >150 µm size fraction and 0–1 cm sediment interval) calculated for the 31 stations analysed in this study (presented from West to East).

Station	Environmental parameters						Foraminiferal parameters		
	Water depth (m)	%Organic matter	Grain size fraction of the sediment					%Tolerant species	%Sensitive species
			%<63 µm	%63–125 µm	%125–250 µm	%250–500 µm	%>500 µm		
Cerb	26	2.20	4.56	0.47	26.30	55.53	13.14	1.90	44.76
Colli	23	1.37	2.89	13.86	39.90	33.74	9.62	2.67	62.03
Leuc	22	1.72	19.73	53.10	24.61	2.41	0.14	46.28	20.92
Gruis	21.5	2.30	42.36	35.64	20.34	1.18	0.49	11.11	24.84
AgdW	18	1.36	18.70	39.71	40.63	0.97	0.00	2.21	22.55
AgdE	21	1.57	8.45	27.56	56.99	6.99	0.00	2.42	34.78
Sete	20	2.30	30.35	31.39	30.66	5.67	1.93	6.54	21.25
Grau	15	3.28	67.92	25.86	6.21	0.00	0.00	7.07	60.51
Bduc	14	1.68	87.52	10.75	1.73	0.00	0.00	17.79	18.40
Fara	10	1.08	6.16	52.35	40.01	1.49	0.00	0.00	27.27
Cart	10	5.91	80.80	13.07	6.14	0.00	0.00	35.06	16.02
Fos	20.8	3.10	73.57	12.42	8.91	4.49	0.61	24.93	23.84
Carry	48	3.54	26.28	14.25	17.57	17.72	24.18	34.77	31.90
Mrade	35	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	16.13	45.16
Mjet	41	5.41	51.91	19.80	19.78	7.61	0.90	39.84	33.13
Plane	40	3.34	13.18	11.24	23.55	26.40	25.63	9.26	57.41
Maire	40	3.31	4.82	4.26	11.22	23.28	56.42	11.11	57.94
Ccan	43	n.d.	13.25	39.96	35.56	7.25	3.97	0.36	52.14
Embi	32	1.87	28.30	50.41	16.44	1.98	2.87	5.10	38.22
Toul	43	7.52	50.90	11.59	4.84	2.82	29.85	22.47	33.33
Porq	40	2.84	10.83	3.24	4.78	13.83	67.31	0.00	76.19
Lav	40	3.78	25.56	13.16	19.98	25.89	15.41	2.33	62.79
Levan	47	2.83	6.47	6.09	13.60	37.84	35.99	5.08	54.24
Pamp	42	2.78	19.10	6.06	11.39	23.95	39.49	8.85	46.15
Fréj	33	4.04	17.35	22.82	33.57	20.25	6.02	2.82	68.75
Antib2	25	4.13	36.56	26.20	16.04	12.91	8.29	0.78	54.09
AntibN	19	1.04	1.62	22.75	59.25	16.37	0.00	0.00	37.39
Nice	30	2.21	46.12	28.47	17.79	6.80	0.81	3.00	16.10
Vfran	42	3.99	14.66	12.25	18.11	22.75	32.22	6.93	51.80
Monac	69	3.72	56.83	20.53	11.39	4.92	6.33	19.51	18.76
Ment	51	1.73	28.26	37.50	32.75	1.49	0.00	21.96	32.80

Station	Foraminiferal parameters								
	%Epiphytic species	%Perforate species	%Porcelaneous species	%Agglutinated species	Absolute densities	Specific richness	Shannon index	Equitability index	ES ₅₀
Cerb	14.29	28.57	29.52	41.90	132	32	3.02	0.87	21.37
Colli	1.60	8.56	57.75	33.69	237	26	1.96	0.60	12.29
Leuc	0.89	46.81	18.62	34.57	709	36	2.71	0.76	16.52
Gruis	1.63	18.30	20.92	60.78	301	24	2.26	0.71	13.33
AgdW	8.58	28.92	13.73	57.35	516	32	2.56	0.74	15.92
AgdE	0.00	11.11	34.78	54.11	260	20	1.95	0.65	11.00
Sete	4.63	28.61	14.99	56.40	461	37	2.59	0.72	16.88
Grau	0.06	78.74	3.02	18.24	2091	33	1.86	0.53	11.55
Bduc	0.00	49.08	2.45	48.47	209	21	2.62	0.86	15.29
Fara	0.00	31.82	27.27	40.91	22	5	1.34	0.83	n.d.
Cart	1.30	49.35	14.29	36.36	287	28	2.51	0.75	14.28
Fos	0.55	67.67	4.93	27.40	459	39	2.83	0.77	17.98
Carry	22.99	62.36	7.76	29.89	439	46	3.07	0.80	20.67
Mrade	24.19	61.29	14.52	24.19	81	23	3.01	0.96	20.65
Mjet	20.73	68.50	7.52	23.98	620	49	3.14	0.81	21.58
Plane	38.89	60.19	13.89	25.93	137	32	3.13	0.90	21.82
Maire	26.19	46.03	30.16	23.81	159	39	3.27	0.89	24.30
Ccan	7.14	20.00	40.71	39.29	354	38	3.02	0.83	19.66
Embi	8.92	24.84	22.29	52.87	199	32	2.93	0.84	19.38
Toul	16.85	52.43	12.36	35.21	337	48	3.46	0.89	26.05
Porq	54.76	71.43	19.05	9.52	51	19	2.48	0.84	18.75
Lav	27.91	50.00	28.57	21.43	43	26	3.10	0.95	n.d.
Levan	12.71	35.59	34.75	29.66	149	34	2.97	0.84	21.81
Pamp	18.85	36.43	25.19	38.37	331	39	2.85	0.78	19.52
Fréj	16.53	33.67	42.94	23.39	627	62	3.52	0.85	25.72
Antib2	16.93	28.79	29.96	41.25	648	56	3.21	0.80	22.88
AntibN	0.00	18.49	36.97	44.54	302	26	2.45	0.75	15.22
Nice	0.00	13.48	15.73	70.79	338	32	2.22	0.64	14.74
Vfran	26.98	41.99	20.06	37.95	874	52	3.41	0.86	24.19
Monac	6.57	41.09	10.32	48.59	664	73	3.71	0.87	27.98
Ment	8.99	36.62	21.56	41.82	486	61	3.48	0.85	25.59

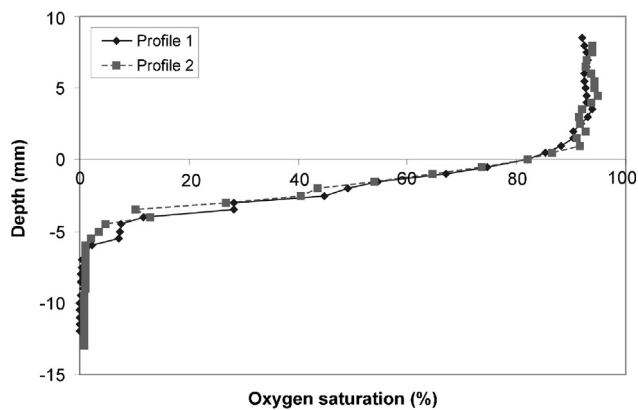
Appendix C.

Linear correlations between environmental and faunal parameters (upper right triangle shows p values and lower left triangle shows r values) considering foraminiferal faunas from the $>150\text{ }\mu\text{m}$ size fraction and 0–1 cm sediment interval. The statistically significant correlations ($p < 0.05$) are indicated in bold.

	Environmental parameters						Foraminiferal parameters												
	Water depth	%OM	%<63 μm	%63–125 μm	%125–250 μm	%250–500 μm	%>500 μm	%Tolerant sp.	%Sensitive sp.	%Epiphytic sp.	%Perforate sp.	%Porcelaneous sp.	%Agglutinated sp.	Absolute densities	Specific richness	Shannon index	Equitability index	ES ₅₀	%Etstd
Environmental parameters																			
Waterdepth		0.07	0.42	0.08	0.15	0.16	0.01	0.45	0.22	0.00	0.27	0.80	0.22	0.72	0.00	0.00	0.00	0.00	0.21
%OM	0.34		0.02	0.02	0.00	0.83	0.18	0.03	0.59	0.07	0.01	0.11	0.03	0.39	0.01	0.00	0.10	0.01	0.05
%<63 μm	–0.16	0.42		0.87	0.00	0.00	0.04	0.01	0.01	0.07	0.02	0.00	0.66	0.06	0.39	0.80	0.47	0.51	0.17
%63–125 μm	–0.33	–0.42	–0.03		0.03	0.00	0.00	0.75	0.02	0.01	0.08	0.85	0.01	0.36	0.61	0.07	0.15	0.22	0.60
%125–250 μm	–0.27	–0.56	–0.60	0.41		0.61	0.03	0.07	0.84	0.11	0.00	0.00	0.09	0.40	0.28	0.08	0.18	0.17	0.28
%250–500 μm	0.27	–0.04	–0.59	–0.62	0.10		0.01	0.06	0.00	0.03	0.46	0.00	0.10	0.11	0.98	0.22	0.16	0.22	0.21
%>500 μm	0.50	0.25	–0.39	–0.63	–0.39	0.46		0.39	0.00	0.00	0.10	0.54	0.00	0.16	0.93	0.11	0.03	0.04	0.67
Foraminiferal parameters																			
%Tolerant sp.	0.15	0.40	0.48	0.06	–0.34	–0.35	–0.17		0.01	0.55	0.01	0.00	0.37	0.40	0.10	0.14	0.68	0.55	0.00
%Sensitive sp.	0.23	0.10	–0.49	–0.44	–0.04	0.59	0.60	–0.48		0.00	0.33	0.01	0.00	0.78	0.92	0.44	0.39	0.28	0.03
%Epiphytic sp.	0.53	0.34	–0.34	–0.50	–0.30	0.40	0.80	–0.12	0.65		0.01	0.95	0.00	0.24	0.48	0.01	0.00	0.01	0.92
%Perforate sp.	0.21	0.50	0.42	–0.33	–0.63	–0.14	0.32	0.48	0.19	0.46		0.00	0.00	0.06	0.44	0.19	0.19	0.25	0.03
%Porcelaneous sp.	–0.05	–0.30	–0.71	–0.04	0.57	0.55	0.12	–0.52	0.50	–0.01	–0.65		0.88	0.06	0.51	0.54	0.79	0.89	0.04
%Agglutinated sp.	–0.24	–0.40	0.09	0.47	0.32	–0.31	–0.52	–0.17	–0.69	–0.60	–0.74	–0.03		0.43	0.67	0.25	0.13	0.16	0.30
Absolutedensities	–0.07	0.17	0.36	0.18	–0.16	–0.30	–0.27	0.16	0.05	–0.22	0.35	–0.35	–0.15		0.04	0.95	0.01	0.63	0.61
Specificrichness	0.66	0.46	0.17	–0.10	–0.21	0.01	0.02	0.31	0.02	0.14	0.15	–0.13	–0.08	0.38		0.00	0.16	0.00	0.07
Shannon index	0.74	0.52	0.05	–0.34	–0.33	0.23	0.30	0.28	0.15	0.45	0.25	–0.12	–0.22	–0.01	0.82		0.00	0.00	0.09
Equitability index	0.52	0.31	–0.14	–0.28	–0.26	0.27	0.42	0.08	0.17	0.57	0.25	–0.05	–0.29	–0.48	0.27	0.69		0.00	0.52
ES ₅₀	0.79	0.47	–0.13	–0.24	–0.27	0.24	0.39	0.12	0.21	0.51	0.23	–0.03	–0.28	–0.10	0.81	0.97	0.82		0.33
%Etstd	0.24	0.36	0.26	0.10	–0.21	–0.24	–0.08	0.97	–0.40	–0.02	0.41	–0.38	–0.20	0.10	0.34	0.32	0.12	0.193	

Appendix D.

Example of an oxygen profile, measured at station Carteau.



Appendix E.

Wilcoxon tests (Z) results and their corresponding probabilities (p) in order to test similarities of major species (>5%) between intervals 0–1 and 0–4 cm.

Stations	n	Z	p
Grau du Roi	18	0.54	0.59
Carteau	18	0.28	0.78
Agde Est	16	0.78	0.44
Pampelone	24	0.06	0.95
Nice	14	0.60	0.55
Ile Maire	21	0.26	0.79
Villefranche	28	0.18	0.86
Menton	29	0.18	0.85
Collioure	17	1.35	0.18
Beauduc	17	0.69	0.49
Toulon Grande Rade	28	0.87	0.39
Carry	30	0.63	0.53
Leucate	24	0.14	0.89
Marseille Jetée	30	0.63	0.53

Appendix F.

Linear correlation (r) between the relative densities of the major species (see Table 2 for the meaning of species abbreviations) and the environmental parameters available for this study. The statistically significant correlations ($p < 0.05$) are indicated in bold.

	Water depth	%Organic matter	Grain size fraction (%)				
			<63 μm	63–125 μm	125–250 μm	250–500 μm	>500 μm
Abecc	-0.52	-0.10	0.54	0.05	-0.11	-0.38	-0.40
Amami	0.31	0.16	-0.24	-0.41	-0.28	0.24	0.69
Astel	0.41	0.32	-0.21	-0.48	-0.35	0.29	0.74
Bgran	-0.38	-0.39	-0.26	0.41	0.39	-0.19	-0.15
Bacul	0.40	-0.13	0.24	0.26	-0.02	-0.30	-0.28
Cauri	0.53	0.56	0.04	-0.29	-0.28	0.08	0.34
Cloba	0.46	0.28	-0.31	-0.38	-0.16	0.52	0.46
Ecris	-0.23	0.05	0.30	0.04	-0.20	-0.17	-0.15
Egran	-0.30	-0.22	0.48	0.06	-0.28	-0.30	-0.25
Epoey	-0.18	0.01	0.50	-0.12	-0.25	-0.21	-0.22
Hboue	0.49	0.51	-0.08	-0.30	-0.18	0.19	0.35
Nterq	0.28	0.16	-0.20	-0.36	-0.24	0.15	0.65
Ndepres	-0.20	-0.21	-0.08	0.44	0.07	-0.15	-0.19
Nscap	-0.18	0.14	0.61	-0.18	-0.34	-0.27	-0.21
Nturg	-0.25	-0.18	0.06	0.40	0.00	-0.24	-0.22
Pmedit	0.35	0.18	-0.21	-0.15	0.04	0.27	0.17
Rphle	0.24	0.59	0.31	-0.16	-0.23	-0.12	-0.01
Rglob	0.35	0.23	-0.34	-0.48	-0.29	0.40	0.77
Spiril	0.23	0.21	-0.26	-0.24	-0.11	0.37	0.36
Vbrad	-0.06	0.47	0.50	-0.11	-0.27	-0.28	-0.16
Along	-0.04	0.26	0.15	-0.24	-0.21	0.19	0.02
Birreg	0.12	-0.15	-0.26	-0.07	0.21	0.27	0.04
Qasp	-0.35	-0.45	-0.48	0.44	0.64	0.02	-0.24
Qbosc	-0.13	-0.18	-0.26	-0.05	0.43	0.14	-0.06
Qcost	0.46	0.06	-0.38	-0.43	-0.17	0.54	0.58
Qsemi	0.19	0.05	-0.08	-0.10	-0.03	0.13	0.11
Sgrata	0.20	0.13	-0.17	-0.20	-0.11	0.22	0.30
Ttrigo	-0.19	-0.29	-0.30	0.01	0.40	0.26	-0.13
Apseudo	0.15	0.52	0.44	-0.22	-0.32	-0.17	-0.03
Escab	-0.58	-0.53	-0.16	0.54	0.60	-0.26	-0.51
LagenamA	-0.03	-0.08	-0.03	0.06	0.04	0.16	-0.16
LagenamB	-0.14	-0.11	0.01	0.23	-0.02	-0.10	-0.10
Pfusc	-0.28	-0.36	0.29	0.28	-0.02	-0.38	-0.32
Rfusif	0.18	0.01	-0.11	-0.29	-0.18	0.25	0.33
Rmica	0.28	0.16	0.12	-0.05	-0.19	-0.10	0.12
Rscorp	0.31	0.37	0.07	-0.12	-0.17	-0.04	0.16
Rsubfus	0.19	0.31	0.18	-0.13	-0.27	-0.03	0.10
Rscot	-0.12	-0.15	-0.09	0.41	0.04	-0.14	-0.13
Taggl	0.32	-0.13	-0.08	0.21	0.14	-0.07	-0.11
Tsagit	0.62	0.37	-0.03	-0.26	-0.23	0.10	0.36

Appendix G.

Evidence from the literature that support our choice to attribute species to stress-tolerant and sensitive (including epiphytic species) groups.

Tolerant species group

Bulimina species (e.g. *B. marginata*, *B. aculeata*, *B. denudata*) are typical of environments with high food input (De Rijk et al., 2000; Morigi et al., 2001; Donnici and Serandrei-Barbero, 2002; Mendes et al., 2004; Eberwein and Mackensen, 2006). For example, *B. marginata* responds to seasonal fluxes of phytodetritus in the Bay of Biscay by increasing its density (Langezaal et al., 2006). *Bulimina* spp. have also been considered as good markers of oxygen-poor conditions (Sen Gupta and Machain-Castillo, 1993; Ohga and Kitazato, 1997; Bernhard and Sen Gupta, 1999; van der Zwaan et al., 1999). ***Cancris auriculus*** and ***Rectuvigerina phlegeri*** are often found in the same assemblages. These species are indicative of eutrophic conditions and stress due to hypoxia (Corliss, 1985; Sen Gupta and Machain-Castillo, 1993; Schmiedl et al., 2000; Milker et al., 2009). More precisely, Diz et al. (2006) described *Rectuvigerina phlegeri* as an opportunistic species rapidly developing after labile organic matter inputs. ***Nonion scaphum*** and ***Nonion depressulum*** are species living in fine sediment with high organic matter inputs (Vénec-Peyré, 1984; Mathieu, 1986; Murray, 1991; Debenay and Redois, 1997; Fontanier et al., 2002; Mojtahid et al., 2006). ***Nonionella turgida***, ***Nonionella stella*** and ***Pseudoeponides falsobecarii*** are all characteristic of fine-grained sediments with high organic matter contents and would be tolerant or even slightly favoured by stressed conditions such as hypoxia (Vénec-Peyré, 1984; Jorissen, 1987; Bernhard and Reimers, 1991; van der Zwaan and Jorissen, 1991; Barmawidjaja et al., 1992; Bernhard et al., 1997; Duijnsteet et al., 2003; Diz et al., 2006). ***Valvulineria bradyana*** is considered as an excellent indicator of sediment enriched in organic matter where environmental stress conditions, such as hypoxia, occur periodically (Jorissen, 1987, 1988; Fontanier et al., 2002). Finally, ***Leptohalysis scottii*** is considered as a strongly opportunistic species because it responds quickly to labile organic matter inputs in the first centimetre of sediment (Scott et al., 2005; Diz and Francés, 2008; Sabbatini et al., 2012). It can support highly turbid waters (Scott et al., 2005; Mojtahid et al., 2009; Goineau et al., 2011) but would only be weakly tolerant to severe hypoxia (Moodley et al., 1997; Ernst et al., 2002; Duijnsteet et al., 2003).

Sensitive species group

According to the literature, **porcelaneous foraminifera** live preferentially in sandy, well oxygenated sediments with relatively low organic matter content (Bizon and Bizon, 1984; Jorissen, 1988; Donnici and Serandrei-Barbero, 2002; Schmiedl et al., 2003). Most of the porcelaneous species will therefore be absent from the assemblage in case of a muddy sediment enriched in organic matter (naturally or anthropogenetically).

The group of **epiphytic species** as described in the main text is sensitive to low oxygen conditions. High percentages suggests the presence of seagrass or macro-algae meadow in the vicinity (Pujos, 1976; Spindler, 1980; Bizon and Bizon, 1984; Jorissen, 1987; Murray, 1991; Langer, 1993; Coppa and Di Tuoro, 1995; Guimerans and Currado, 1999; van der Zwaan et al., 1999; Villanueva Guimerans and Cervera Currado, 1999; Mendes et al., 2004; Panieri et al., 2005; Mojtahid et al., 2006; Schönfeld, 2002; Martins et al., 2007; Milker et al., 2009). In addition to the sessile and temporarily motile species from morphotypes A and B (Langer, 1993) considered in our “epiphytic group”, we added some species from the motile epiphytic morphotypes C and D in the sensitive

species. This concerns *Reussella spinulosa*, *Spirillina* and *Elphidium* species. According to a study in the Adriatic Sea, *Reussella spinulosa* would show a certain preference for a sandy substratum with a low input of clay (Jorissen, 1987). *Elphidium crispum* shows no specific preference to a particular type of sediment. In the study of Jorissen (1987), this species is found at sites where the organic matter content is slightly elevated but it is very rare in stations under the direct influence of the Po river output. This species is also considered as a motile epiphytic suspension feeder (Langer, 1993). Therefore, this species would not support severe stress conditions. In our material, *Elphidium granosum* and *Elphidium poeyanum* are mainly represented by the *lidoense* and *decipiens* morphotypes, respectively. These two morphotypes, which have been considered as sensitive by Jorissen (1987), are mainly found in silty to sandy areas with a relatively low organic matter content, probably with well oxygenated bottom waters.

Supplementary data

Supplementary data associated with this article can be found, in the online version, at <http://dx.doi.org/10.1016/j.ecolind.2013.09.028>.

References

- Aloisi, J.C., Got, H., Monaco, A., 1973. Carte géologique du précontinent languedocien du Cap Leucate à la pointe de Beauduc au 1/250000ième. International Institute for Aerial survey and Earth science (ITC) (Editor), Netherlands.
- Altenbach, A.V., Sarnthein, M., 1989. Productivity record in benthic foraminifera. In: Berger, W.H., Smetacek, V.S., Wefer, G. (Eds.), Productivity of the Ocean: Present and Past. Wiley, Chichester, pp. 255–269.
- Alve, E., 1991. Benthic foraminifera in sediment cores reflecting heavy-metal pollution in Sørkjord, Western Norway. Journal of Foraminiferal Research 21 (1), 1–19.
- Alve, E., 1995. Benthic foraminiferal responses to estuarine pollution: a review. Journal of Foraminiferal Research 25, 190–203.
- Alve, E., Nagy, J., 1986. Estuarine foraminiferal distribution in Sandebukta, a branch of the Oslo Fjord. Journal of Foraminiferal Research 16, 157–175.
- Armynot du Chatelet, E., Debenay, J.P., Soullard, R., 2004. Foraminiferal proxies for pollution monitoring in moderately polluted harbors. Environmental Pollution 127, 27–40.
- Bandy, O.L., Ingle, J.C., Resig, J.M., 1964. Foraminifera, Los Angeles County outfall area, California. Limnology and Oceanography 9, 124–137.
- Bandy, O.L., Ingle, J.C., Resig, J.M., 1965. Modification of foraminiferal distribution by the Orange County outfall, California. Ocean Science and Ocean Engineering, Marine Technology Society, Transactions, 54–76.
- Barmawidjaja, D.M., Jorissen, F.J., Puskaric, S., van der Zwaan, G.J., 1992. Microhabitat selection by benthic foraminifera in the northern Adriatic Sea. Journal of Foraminiferal Research 22 (4), 297–317.
- Barmawidjaja, D.M., van der Zwaan, G.J., Jorissen, F.J., Puskaric, S., 1995. 150 years of eutrophication in the northern Adriatic Sea, evidence from a benthic foraminiferal record. Marine Geology 122, 367–384.
- Bassetti, M.A., et al., 2006. Sand bodies at the shelf edge in the Gulf of Lions (Western Mediterranean): deglacial history and modern processes. Marine Geology 234 (1–4), 93–109.
- Bergamin, L., et al., 2009. Benthic foraminifera from the coastal zone of Baia (Naples, Italy): assemblage distribution and modification as tools for environmental characterisation. Marine Pollution Bulletin 59 (8–12), 234–244.
- Berné, S., Gorini, C., 2005. The Gulf of Lions: an overview of recent studies within the French ‘Margins’ program. Marine and Petroleum Geology 22, 691–693.
- Bernhard, J.M., 1988. Postmortem vital staining in benthic foraminifera: duration and importance in population and distributional studies. Journal of Foraminiferal Research 18, 143–146.
- Bernhard, J.M., Reimers, C.E., 1991. Benthic foraminiferal population fluctuations related to anoxia: Santa Barbara Basin. Biogeochemistry 15 (2), 1577–1585.
- Bernhard, J.M., Sen Gupta, B.K., 1999. Foraminifera of oxygen-depleted environments. In: Sen Gupta, B.K. (Ed.), Modern Foraminifera. Kluwer Academic Press, Dordrecht.
- Bernhard, J.M., Sen Gupta, B.K., Borne, P.F., 1997. Benthic foraminiferal proxy to estimate dysoxic bottom-water oxygen concentrations: Santa Barbara basin, US Pacific continental margin. Journal of Foraminiferal Research 27 (4), 301–310.
- Bizon, G., Bizon, J.J., 1984. Distribution des foraminifères sur le plateau continental au large du Rhône. In: Bizon, J.J., Burrollet, P.F. (Eds.), Ecologie des Microorganismes en Méditerranée occidentale ‘ECOMED’. Association Française des Techniciens du Pétrole (AFTP), Paris, pp. 84–94.
- Blanc-Vernet, L., 1969. Contribution à l’étude des foraminifères de Méditerranée. Recueil des Travaux de la Station Marine d’Endoume 64 (48), 1–135.

- Blanc-Vernet, L., 1984. Les foraminifères de l'herbier de *Posidonia oceanica* en Méditerranée: analyse des assemblages, aspects régionaux, application aux microfaunes fossiles. In: Boudouresque, C.F., Jeudy de Grissac, A., Olivier, J. (Eds.), International Workshop on Posidonia Oceanica Beds. GIS Posidonie Publications, Marseille, pp. 3–14.
- Borja, A., Franco, J., Pérez, V., 2000. A marine biotic index to establish the ecological quality of soft-bottom benthos within European estuarine and coastal environments. *Marine Pollution Bulletin* 40 (12), 1100–1114.
- Bosc, E., Bricaud, A., Antoine, D., 2004. Seasonal and interannual variability in algal biomass and primary production in the Mediterranean Sea, as derived from 4 years of SeaWiFS observations. *Global Biogeochemical Cycles* 18 (1.).
- Bouchet, V.M.P., Alve, E., Rygg, B., Telford, R.J., 2012. Benthic foraminifera provide a promising tool for ecological quality assessment of marine waters. *Ecological Indicators* 23, 66–75.
- Bouchet, V.M.P., Debenay, J.P., Sauriau, P.-G., Radford-Knoery, J., Soletchnik, P., 2007. Effects of short-term environmental disturbances on living benthic foraminifera during the Pacific oyster summer mortality in the Marennes-Oléron Bay (France). *Marine Environmental Research* 64, 358–383.
- Boudouresque, C.F., et al., 2006. Préservation et conservation des herbiers à *Posidonia oceanica*. Ramoge Pub, pp. 1–202.
- Boudouresque, C.F., et al., 2000. A monitoring network based on the seagrass *Posidonia oceanica* in the Northwestern Mediterranean Sea. *Biologia Marina Mediterranea* 7 (2), 328–331.
- Buosi, C., Arminot du Chatelet, E., Cherchi, A., 2012. Benthic foraminiferal assemblages in the current-dominated Strait of Bonifacio (Mediterranean Sea). *Journal of Foraminiferal Research* 42 (1), 39–55.
- Burone, L., Venturini, N., Sprechmann, P., Valente, P., Muniz, P., 2006. Foraminiferal responses to polluted sediments in the Montevideo coastal zone, Uruguay. *Marine Pollution Bulletin* 52 (1), 61–73.
- Buzas, M.A., Culver, S.J., Jorissen, F.J., 1993. A statistical evaluation of the microhabitats of living (stained) infaunal benthic foraminifera. *Marine Micropaleontology* 20 (3–4), 311–320.
- Cherchi, A., et al., 2009. Benthic foraminifera response and geochemical characterization of the coastal environment surrounding the polluted industrial area of Portovesme (South-Western Sardinia, Italy). *Marine Pollution Bulletin* 59 (8–12), 281–296.
- Clark, D.E., 1971. Effects of aquaculture outfall on benthonic foraminifera in Clam Bay, Nova Scotia. *Maritime Sediments* 7, 76–84.
- Coccioni, R., Frontalini, F., Marsili, A., Mana, D., 2009. Benthic foraminifera and trace element distribution: a case-study from the heavily polluted lagoon of Venice (Italy). *Marine Pollution Bulletin* 59 (8–12), 257–267.
- Coppa, M.G., Di Tuoro, A., 1995. Preliminary data on the Holocene foraminifera of the Cilento continental shelf (Tyrrhenian Sea). *Revista Espanola de Paleontologia* 10 (2), 161–174.
- Corliss, B.H., 1985. Microhabitats of benthic foraminifera within deep-sea sediments. *Nature* 314, 435–438.
- Corliss, B.H., 1991. Morphology and microhabitat preferences of benthic foraminifera from the northwest Atlantic Ocean. *Marine Micropaleontology* 17 (3–4), 195–236.
- Corliss, B.H., Emerson, S., 1990. Distribution of Rose Bengal stained deep-sea benthic foraminifera from the Nova Scotian continental margin and Gulf of Maine. *Deep-Sea Research Part I* 37 (3), 381–400.
- Dauvin, J.C., Ruellet, T., 2007. Polychaete/amphipod ratio revisited. *Marine Pollution Bulletin* 55, 215–224.
- de Noijer, L.J., Duijnste, I.A.P., Bergman, M.J.N., van der Zwaan, G.J., 2008. The ecology of benthic foraminifera across the Frisian Front, southern North Sea. *Estuarine, Coastal and Shelf Science* 78 (4), 715–726.
- De Rijk, S., Jorissen, F.J., Rohling, E.J., Troelstra, S.R., 2000. Organic flux control on bathymetric zonation of Mediterranean benthic foraminifera. *Marine Micropaleontology* 40 (3), 151–166.
- Debenay, J.P., 2000. Foraminifera of tropical paralic environments. *Micropaleontology* 46 (Supplement 1), 153–160.
- Debenay, J.-P., Redois, F., 1997. Distribution of the twenty seven dominant species of shelf benthic foraminifera on the continental shelf, north of Dakar (Senegal). *Marine Micropaleontology* 29 (3–4), 237–255.
- Denoyelle, M., Jorissen, F., Martin, D., Galgani, F., Miné, J., 2010. Comparison of benthic foraminifera and macrofaunal indicators of the impact of oil-based drill mud disposal. *Marine Pollution Bulletin* 60 (11), 2007–2021.
- Devlin, M., Best, M., Haynes, D., 2007. Implementation of the Water Framework Directive in European marine waters. *Marine Pollution Bulletin* 55 (Spec. Issue 1–6), 1–6.
- Diaz, R.J., Solan, M., Valente, R.M., 2004. A review of approaches for classifying benthic habitats and evaluating habitat quality. *Journal of Environmental Management* 73, 165–181.
- Diz, P., Francés, G., 2008. Distribution of live benthic foraminifera in the Ría de Vigo (NW Spain). *Marine Micropaleontology* 66 (3–4), 165–191.
- Diz, P., Francés, G., Rosón, G., 2006. Effects of contrasting upwelling–downwelling on benthic foraminiferal distribution in the Ría de Vigo (NW Spain). *Journal of Marine Systems* 60 (1–2), 1–18.
- Donnici, S., Serandrei-Barbero, R., 2002. The benthic foraminiferal communities of the northern Adriatic continental shelf. *Marine Micropaleontology* 44 (3–4), 93–123.
- Duchemin, G., et al., 2008. New monitoring tool for assessing environmental impact of off-shore drilling activities: *Benthic foraminifera*. Society of Petroleum Engineers Journal, Paper 111959, 8 pp.
- Duijnste, I., de Lugt, I., Vonk Noordegraaf, H., van der Zwaan, B., 2004. Temporal variability of foraminiferal densities in the northern Adriatic Sea. *Marine Micropaleontology* 50 (1–2), 125–148.
- Duijnste, I.A.P., Ernst, S.R., van der Zwaan, G.J., 2003. Effect of anoxia on the vertical migration of benthic foraminifera. *Marine Ecology Progress Series* 246, 85–94.
- Duros, P., et al., 2011. Live (stained) benthic foraminifera in the Whittard Canyon, Celtic margin (NE Atlantic). *Deep-Sea Research Part I* 58 (2), 128–146.
- Durrieu de Madron, X., et al., 2000. Particulate matter and organic carbon budgets for the Gulf of Lions (NW Mediterranean). *Oceanologica Acta* 23 (6), 717–730.
- Durrieu, J., et al., 2006. Aged drilled cuttings offshore Gabon: new methodology for assessing their impact. Society of Petroleum Engineers Journal, Paper 98414: 8 pp.
- Eberwein, A., Mackensen, A., 2006. Regional primary productivity differences off Morocco (NW-Africa) recorded by modern benthic foraminifera and their stable carbon isotopic composition. *Deep Sea Research I* 53 (8), 1379–1405.
- Ernst, S., Bours, R., Duijnste, I., van der Zwaan, B.D., 2005. Experimental effects of an organic matter pulse and oxygen depletion on a benthic foraminiferal shelf community. *Journal of Foraminiferal Research* 35 (3), 177–197.
- Ernst, S., Duijnste, I., van der Zwaan, B., 2002. The dynamics of the benthic foraminiferal microhabitat: recovery after experimental disturbance. *Marine Micropaleontology* 46 (3–4), 343–361.
- Fontanier, C., et al., 2008. Live foraminifera from the open slope between Grand Rhône and Petit Rhône Canyons (Gulf of Lions, NW Mediterranean). *Deep-Sea Research Part I* 55 (11), 1532–1553.
- Fontanier, C., et al., 2002. Live benthic foraminiferal faunas from the Bay of Biscay: faunal density, composition, and microhabitats. *Deep-Sea Research Part I* 49, 751–785.
- Frontalini, F., et al., 2009. Benthic foraminifera as bio-indicators of trace element pollution in the heavily contaminated Santa Gilla lagoon (Cagliari, Italy). *Marine Pollution Bulletin* 58 (6), 858–877.
- Frontalini, F., Coccioni, R., 2008. Benthic foraminifera for heavy metal pollution monitoring: a case study from the central Adriatic Sea coast of Italy. *Estuarine, Coastal and Shelf Science* 76, 404–417.
- Frontalini, F., Coccioni, R., 2011. Benthic foraminifera as bioindicators of pollution: a review of Italian research over the last three decades. *Revue de micropaléontologie* 54, 115–127.
- Gobert, S., et al., 2009. Assessment of the ecological status of Mediterranean French coastal waters as required by the Water Framework Directive using the *Posidonia oceanica* Rapid Easy Index: PREI. *Marine Pollution Bulletin* 58 (11), 1727–1733.
- Godshalk, G.L., Wetzel, R.G., 1978. Decomposition of aquatic angiosperms III. *Zostera marina* L. and a conceptual model of decomposition. *Aquatic Botany* 5, 329–354.
- Goineau, A., et al., 2012. Temporal variability of live (stained) benthic foraminiferal faunas in a river-dominated shelf? Faunal response to rapid changes of the river influence (Rhône prodelta, NW Mediterranean). *Biogeosciences* 9, 1367–1388.
- Goineau, A., et al., 2011. Live (stained) benthic foraminifera from the Rhône prodelta (Gulf of Lion, NW Mediterranean): environmental controls on a river-dominated shelf. *Journal of Sea Research* 65 (1), 58–75.
- Gomez Gesteira, L., Dauvin, J.C., 2000. Amphipods are good bioindicators of the impact of oil spills on soft-bottom macrobenthic communities. *Marine Pollution Bulletin* 40, 1017–1027.
- Goody, A., Levin, L.A., Linke, P., Heeger, T., 1992. The role of benthic foraminifera in deep-sea food webs and carbon cycling. In: G.T.a.P Rowe, V., Rowe, G.T., Pariente, V. (Eds.), *Deep-Sea Food Chains and the Global Carbon cycle*. Kluwer Academic Publishers, The Netherlands, pp. 63–91.
- Goody, A.J., 1994. The biology of deep-sea foraminifera: a review of some advantages and their applications in paleoceanography. *Palaios* 9, 14–31.
- Goody, A.J., Bernhard, J.M., Levin, L.A., Suhr, S.B., 2000. Foraminifera in the Arabian Sea oxygen minimum zone and other oxygen-deficient settings: taxonomic composition, diversity, and relation to metazoan faunas. *Deep-Sea Research Part II* 47 (1–2), 25–54.
- Goody, A.J., Rathburn, A.E., 1999. Temporal variability in living deep-sea foraminifera: a review. *Earth-Science Reviews* 46, 187–212.
- Guimerans, P.V., Currado, J.L.C., 1999. Distribution of Planorbulinacea (benthic foraminifera) assemblages in surface sediments on the northern margin of the Gulf of Cadiz. *Boletín del Instituto Español de Oceanografía* 15 (1–4), 181–190.
- Hammer, Ø., Harper, D., 2005. *Paleontological Data Analysis*. Blackwell Oxford, pp. 351.
- Harrison, P.G., 1989. Detrital processing in seagrass systems: a review of factors affecting decay rates, remineralization and detritivory. *Aquatic Botany* 35, 263–288.
- Hayek, L.E.C., Buzas, M.A., 1997. *Surveying Natural Populations*. Columbia University Press, New York, pp. 563.
- Hedges, J.L., Keil, R.G., 1995. Sedimentary organic matter preservation: an assessment and speculative synthesis. *Marine Chemistry* 49, 81–115.
- Helder, W., Bakker, J.F., 1985. Shipboard comparison of micro- and mini-electrodes for measuring oxygen in marine sediments. *Limnology and Oceanography* 30, 1106–1109.
- Herguera, J.C., Berger, W.H., 1991. Paleoproductivity from benthic foraminifera abundance: glacial and postglacial change in the west-equatorial Pacific. *Abstracts with Program Geological Society of America* 23 (5), 107.
- Hess, S., Kuhnt, W., 1996. Deep-sea benthic foraminiferal recolonization of the 1991 Mt. Pinatubo ash layer in the South China Sea. *Marine Micropaleontology* 28, 171–197.
- Hurlbert, S.H., 1971. The nonconcept of species diversity: a critique and alternative parameters. *Ecology* 52 (4), 577–586.

- Hyams-Kaphzan, O., Almogi-Labin, A., Benjamini, C., Herut, B., 2009. Natural oligotrophy vs. pollution-induced eutrophy on the SE Mediterranean shallow shelf (Israel): environmental parameters and benthic foraminifera. *Marine Pollution Bulletin* 58 (12), 1888–1902.
- Jorissen, F., 1988. Benthic foraminifera from the Adriatic Sea; principles of phenotypic variation. *Utrecht Micropaleontological Bulletins* 37, 174.
- Jorissen, F.J., 1987. The distribution of benthic foraminifera in the Adriatic Sea. *Marine Micropaleontology* 12, 21–48.
- Jorissen, F.J., Barmawidjaja, D.M., Puskarić, S., van der Zwaan, G.J., 1992. Vertical distribution of benthic foraminifera in the northern Adriatic Sea: the relation with the organic flux. *Marine Micropaleontology* 19, 131–146.
- Jorissen, F.J., et al., 2009. Impact of oil-based drill mud disposal on benthic foraminiferal assemblages on the continental margin off Angola. *Deep-Sea Research Part II* 56 (23), 2270–2291.
- Jorissen, F.J., de Stigter, H.C., Widmark, J.G.V., 1995. A conceptual model explaining benthic foraminiferal microhabitats. *Marine Micropaleontology* 22, 3–15.
- Jorissen, F.J., Fontanier, C., Thomas, E., 2007. Chapter seven paleoceanographical proxies based on deep-sea benthic foraminiferal assemblage characteristics. In: Hillaire-Marcel, C., De Vernal, A. (Eds.), *Paleoceanography of the Late Cenozoic. Developments in Marine Geology*, Elsevier, pp. 263–325.
- Jorissen, F.J., Wittling, I., Peyrouquet, J.P., Rabouille, C., Relexans, J.C., 1998. Live benthic foraminiferal faunas off Cap Blanc, NW Africa: community structure and microhabitats. *Deep-Sea Research Part I* 45, 2157–2188.
- Kim, J.-H., et al., 2006. Origin and distribution of terrestrial organic matter in the NW Mediterranean (Gulf of Lions): exploring the newly developed BIT index. *Geochemistry Geophysics Geosystems* 7, <http://dx.doi.org/10.1029/2006GC001306>.
- Kitazato, H., 1988. Ecology of benthic foraminifera in the tidal zone of a rocky shore. *Revue de micropaléontologie (Spec. No. 2)*, 815–825.
- Klap, V.A., Hemminga, M.A., Boon, J.J., 2000. Retention of lignin in seagrasses: angiosperms that returned to the sea. *Marine Ecology Progress Series* 194, 1–11.
- Langer, M.R., 1993. Epiphytic Foraminifera. *Marine Micropaleontology* 20 (3–4), 235–265.
- Langezaal, A.M., et al., 2006. The influence of seasonal processes on geochemical profiles and foraminiferal assemblages on the outer shelf of the Bay of Biscay. *Continental Shelf Research* 26 (15), 1730–1755.
- Licari, L.N., Schumacher, S., Wenzhöfer, F., Zabel, M., Mackensen, A., 2003. Communities and microhabitats of living benthic foraminifera from the tropical east Atlantic: impact of different productivity regimes. *Journal of Foraminiferal Research* 33, 10–31.
- Martins, V., et al., 2007. A multiproxy approach of the Holocene evolution of shelf-slope circulation on the NW Iberian Continental Shelf. *Marine Geology* 239, 1–18.
- Mateo, M.A., Cebrán, J., Dunton, K., Mutchler, T., 2006. Carbon flux in seagrass ecosystems. In: Larkum, A.W.D., Orth, R.J., Duarte, C.M. (Eds.), *Seagrasses: biology, ecology and conservation*. Springer, Dordrecht, pp. 159–192.
- Mathieu, R., 1986. Sédiments et foraminifères actuels de la marge continentale atlantique du maroc, n° 86-14. Thèse Sci. Univ. Paris VI, 420 pp.
- Mendes, I., Gonzalez, R., Dias, J.M.A., Lobo, F., Martins, V., 2004. Factors influencing recent benthic foraminifera distribution on the Guadiana shelf (Southwestern Iberia). *Marine Micropaleontology* 51, 171–192.
- Milker, Y., et al., 2009. Distribution of recent benthic foraminifera in shelf carbonate environments of the Western Mediterranean Sea. *Marine Micropaleontology* 73 (3–4), 207–225.
- Millot, C., Taupier-Letage, I., 2005. Additional evidence of LIW entrainment across the Algerian subbasin by mesoscale eddies and not by a permanent westward flow. *Progress in Oceanography* 66 (2–4), 231–250.
- Mojtahid, M., 2007. Les foraminifères benthiques: bio-indicateurs d'eutrophisation naturelle et anthropique en milieu marin franc. Angers, France, pp. 390.
- Mojtahid, M., et al., 2006. Benthic foraminifera as bio-indicators of drill cutting disposal in tropical east Atlantic outer shelf environments. *Marine Micropaleontology* 61 (1–3), 58–75.
- Mojtahid, M., Jorissen, F., Lansard, B., Fontanier, C., 2010. Microhabitat selection of benthic foraminifera in sediments off the Rhône River Mouth (NW Mediterranean). *Journal of Foraminiferal Research* 40 (3), 231–246.
- Mojtahid, M., et al., 2009. Spatial distribution of live benthic foraminifera in the Rhône prodelta: faunal response to a continental–marine organic matter gradient. *Marine Micropaleontology* 70, 177–200.
- Mojtahid, M., Jorissen, F., Pearson, T.H., 2008. Comparison of benthic foraminiferal and macrofaunal responses to organic pollution in the Firth of Clyde (Scotland). *Marine Pollution Bulletin* 56 (1), 42–76.
- Moodley, L., van der Zwaan, G.J., Herman, P.M.J., Kempers, L., Van Breugel, P., 1997. Differential response of benthic meiofauna to anoxia with special reference to foraminifera (Protista: Sarcodina). *Marine Ecology Progress Series* 158, 151–163.
- Morigi, C., Jorissen, F.J., Gervais, A., Guichard, S., Borsetti, A.M., 2001. Benthic foraminiferal faunas in surface sediments off NW Africa: relationship with the organic flux to the ocean floor. *Journal of Foraminiferal Research* 31, 350–368.
- Morvan, J., Le Cadre, V., Jorissen, F., Debenay, J.P., 2004. Foraminifera as potential bio-indicators of the “Erika” oil spill in the Bay of Bourgneuf: field and experimental studies. *Aquatic Living Resources* 17 (3), 317–322.
- Murray, J.W., 1973. *Distribution and Ecology of Living Benthic Foraminiferids*. Heine-mann Educational Books, London, pp. 274.
- Murray, J.W., 1986. Living and dead Holocene foraminifera of Lyme Bay, southern England. *Journal of Foraminiferal Research* 16, 347–352.
- Murray, J.W., 1989. Syndepositional dissolution of calcareous foraminifera in modern shallow water sediments. *Marine micropaleontology* 15, 117–121.
- Murray, J.W., 1991. *Ecology and distribution*. In: BENTHOS'90, Tokai University Press, Sendai.
- Murray, J.W., 2006. *Ecology and Applications of Benthic Foraminifera*. Cambridge University Press, Cambridge, 426 pp.
- Nigam, R., Saraswat, R., Panchang, R., 2006. Application of foraminifera in ecotoxicology: retrospect, prospect and prospect. *Environment International* 32, 273–283.
- Ohga, T., Kitazato, H., 1997. Seasonal changes in bathyal foraminiferal populations in response to the flux of organic matter (Sagami Bay, Japan). *Terra Nova* 9 (1), 33–37.
- Østergaard Pedersen, M., Serrano, O., Mateo, M.A., Holmer, M., 2011. Temperature effects on decomposition of a *Posidonia oceanica* mat. *Aquatic microbial ecology* 65, 169–182.
- Pairaud, I., Desmare, S., 2011. *Courantologie de la sous-région marine Méditerranée occidentale DCSMM/EI/MO*, Ministère de l'Écologie, du Développement Durable, des Transports et du Logement, Ref. DCSMM/EI/EE/MO/1.1.6/2011, p. 10.
- Panieri, G., Gamberi, F., Marani, M., Barbieri, R., 2005. Benthic foraminifera from a recent, shallow-water hydrothermal environment in the Aeolian Arc (Tyrrhenian Sea). *Marine Geology* 218, 207–229.
- Pearson, T.H., Rosenberg, R., 1978. Macrobenthic succession in relation to organic enrichment and pollution of the marine environment. *Oceanography and Marine Biology: An Annual Review* 16, 229–311.
- Pielou, E.C., 1966. The measurement of diversity in different types of biological collections. *Journal of Theoretical Biology* 13, 131–144.
- Pielou, E.C., 1975. *Ecological Diversity*. Wiley, New York, pp. 165.
- Platon, E., Sen Gupta, B.K., Rabalais, N.N., Turner, R.E., 2005. Effect of seasonal hypoxia on the benthic foraminiferal community of the Louisiana inner continental shelf, the 20th century record. *Marine Micropaleontology* 54, 263–283.
- Pont, D., 1997. Les débits solides du Rhône à proximité de son embouchure: données récentes (1994–1995). *Revue de géographie* 72, 13–33.
- Pont, D., Simonnet, J.P., Walter, A.V., 2002. Medium-term changes in suspended sediment delivery to the ocean: consequences of catchment heterogeneity and river management (Rhône river, France). *Estuarine, Coastal and Shelf Science* 54, 1–18.
- Pujos, M., 1976. *Ecologie des foraminifères benthiques et des thecamoebiens de la Gironde et du plateau continental sud-Gasconne. Application à la connaissance du Quaternaire terminal de la région ouest-Gironde. Mémoires de l'Institut de Géologie du Bassin d'Aquitaine*, 314 pp.
- Pusceddu, A., Dell'Anno, A., Fabian, M., Danovaro, R., 2009. Quantity and bioavailability of sediment organic matter as signatures of benthic trophic status. *Marine Ecology Progress Series* 375, 41–52.
- Raimbault, P., Durrieu de Madron, X., 2003. Research activities in the Gulf of Lion (NW Mediterranean) within the 1997–2001 PNEC project. *Oceanologica Acta* 26, 291–298.
- Rathburn, A.E., Corliss, B.H., 1994. The ecology of living (stained) deep-sea benthic foraminifera from the Sulu Sea. *Paleoceanography* 9 (1), 87–150.
- Rathburn, A.E., Corliss, B.H., Tappa, K.D., Lohmann, K.C., 1996. Comparisons of the ecology and stable isotopic compositions of living (stained) benthic foraminifera from the Sulu and South China Seas. *Deep-Sea Research Part I* 43 (10), 1617–1646.
- Redois, F., Debenay, J.-P., 1996. Influence du confinement sur la répartition des foraminifères benthiques: exemple de l'estran d'une ria mésotidale de Bretagne méridionale. *Revue de Paléobiologie* 15 (1), 243–260.
- Resig, J.M., 1960. Foraminiferal ecology around ocean outfalls off Southern California, waste disposal in the marine environment. Pergamon Press, London, pp. 104–121.
- Revsbech, N.P., 1983. In-situ measurements of oxygen profiles of sediments by use of oxygen microelectrodes. In: Ganuger, E.F.H. (Ed.), *Polarographic Oxygen Sensors*. Springer, Berlin, pp. 265–273.
- Revsbech, N.P., Jørgensen, B.B., 1986. Microelectrodes: their use in microbial ecology. *Advances in Microbial Ecology* 9, 293–352.
- Romano, E., et al., 2009. The impact of the Bagnoli industrial site (Naples, Italy) on sea-bottom environment. Chemical and textural features of sediments and the related response of benthic foraminifera. *Marine Pollution Bulletin* 59 (8–12), 245–256.
- Rosenberg, R., Blomqvist, M., Nilsson, H.C., Cederwall, H., Dimming, A., 2004. Marine quality assessment by use of benthic species-abundance distributions: a proposed new protocol within the European Union Water Framework Directive. *Marine Pollution Bulletin* 49, 728–739.
- Sabbatini, A., et al., 2012. Foraminiferal assemblages and trophic state in coastal sediments of the Adriatic Sea. *Journal of Marine Systems* 105, 163–174.
- Sabbatini, A., et al., 2010. Modern benthic foraminifera at Northern shallow sites of Adriatic Sea and soft-walled, monothalamous taxa: a brief overview. *Micropaleontology* 56 (3–4), 359–376.
- Sanders, H.L., 1968. Marine benthic diversity: a comparative study. *American Naturalist* 102 (925), 243–282.
- Schmiedl, G., et al., 2000. Trophic control of benthic foraminiferal abundance and microhabitat in the bathyal Gulf of Lions, western Mediterranean Sea. *Marine Micropaleontology* 40, 167–188.
- Schmiedl, G., et al., 2003. Benthic foraminiferal record of ecosystem variability in the eastern Mediterranean Sea during times of sapropel S-5 and S-6 deposition. *Palaeogeography, Palaeoclimatology, Palaeoecology* 190, 139–164.
- Schönfeld, J., 2002. Recent benthic foraminiferal assemblages in deep high-energy environments from the Gulf of Cadiz (Spain). *Marine Micropaleontology* 44 (3–4), 141–162.

- Schönfeld, J., et al., 2012. The FOBIMO (Foraminiferal Blo-MOnitoring) initiative towards a standardised protocol for soft-bottom benthic foraminiferal monitoring studies. *Marine Micropaleontology* 94–95, 1–13.
- Scott, D.B., et al., 2005. Pollution monitoring in two North American estuaries: historical reconstructions using benthic foraminifera. *Journal of Foraminiferal Research* 35, 65–82.
- Scott, G.A., Scourse, J.D., Austin, W.E.N., 2003. The distribution of benthic foraminifera in the Celtic Sea: the significance of seasonal stratification. *Journal of Foraminiferal Research* 33 (1), 32–61.
- Seiglie, G.A., 1968. Foraminiferal assemblages as indicators of high organic carbon content in sediments and of polluted waters. *American Association of Petroleum Geologists Bulletin* 52, 2231–2241.
- Seiglie, G.A., 1971. A preliminary note on the relationships between foraminifera and pollution in two Puerto Rican bays. *Caribbean Journal of Science* 11, 93–98.
- Sempéré, R., Charrière, B., VanWambeke, F., Cauwet, G., 2000. Carbon inputs of the Rhône River to the Mediterranean Sea. *Global Biogeochemical Cycles* 14, 669–681.
- Sen Gupta, B.K., Machain-Castillo, M.L., 1993. Benthic foraminifera in oxygen-poor habitats. *Marine Micropaleontology* 20 (3–4), 183–201.
- Shirayama, Y., 1984. The abundance of deep sea meiobenthos in the Western Pacific in relation to environmental factors. *Oceanologica Acta* 7, 113–121.
- Simboura, N., Zenetos, A., 2002. Benthic indicators to use in ecological quality classification of Mediterranean soft bottoms marine ecosystems, including a new biotic index. *Mediterranean Marine Science* 3/2, 77–111.
- Spindler, M., 1980. The pelagic golfweed *Sargassum natans* as a habitat for the benthic foraminifera *Planorbulina acervalis* et *Rosalina globularis*. *Neues Jahrbuch für Mineralogie, Geologie und Palaontologie, Monatshefte* 9, 569–580.
- Teodoro, A.C., et al., 2010. Analysis of foraminifera assemblages and sediment geochemical properties to characterise the environment near Araca and Saco da Cape la domestic sewage submarine outfalls of Sao Sebastiao Channel, Sao Paulo State, Brazil. *Marine Pollution Bulletin* 60 (4), 536–553.
- Thill, A., et al., 2001. Evolution of particle size and concentration in the Rhône river mixing zone: influence of salt flocculation. *Continental Shelf Research* 21, 2127–2140.
- Tyson, R.V., 1995. *Sedimentary organic matter. Organic facies and palynofacies*. Chapman & Hall, London, pp. 615.
- van der Zwaan, G.J., et al., 1999. Benthic foraminifera: proxies or problems? A review of paleoecological concepts. *Earth-Sciences Reviews* 46, 213–236.
- van der Zwaan, G.J., Jorissen, F.J., 1991. Biofacial patterns in river-induced shelf anoxia. In: Tyson, R.V., Pearson, T.H. (Eds.), *Modern and Ancient Continental Shelf Anoxia*. Geological Society Special Publication No. 58, London, pp. 65–82.
- Vénec-Peyré, M.T., 1984. Etude de la distribution des foraminifères vivant dans la Baie de Banyuls-sur-Mer. In: Bizon, J.J., Burolet, P.F. (Eds.), *Ecologie des microorganismes en Méditerranée occidentale "Ecomed"*. Association française des Techniciens du Pétrole, Paris, pp. 60–80.
- Vénec-Peyré, M.T., Le Calvez, Y., 1981. Etude des foraminifères de l'herbier à Posidonies de Banyuls-sur-Mer. In: 106th Congrès national des sociétés savantes, Perpignan, pp. 191–203.
- Vénec-Peyré, M.T., Le Calvez, Y., 1988. Les foraminifères épiphytes de l'herbier de Posidonies de Banyuls-sur-Mer (Méditerranée occidentale): Etude des variations spatiotemporelles du peuplement. *Cahiers de micropaleontologie*, NS 3, 21–40.
- Vilela, C.G., Batista, D.S., Baptista Neto, J.A., Ghiselli Jr., R.O., 2011. Benthic foraminifera distribution in a tourist lagoon in Rio de Janeiro, Brazil: a response to anthropogenic impacts. *Marine Pollution Bulletin* 62 (10), 2055–2074.
- Villanueva Guimerans, P., Cervera Currado, J.L., 1999. Distribution of *Planorbulina* (benthic foraminifera) assemblages in surface sediments on the northern margin of Gulf of Cadiz. *Boletín del Instituto Español de Oceanografía* 15, 181–190.
- Walton, W.R., 1952. Techniques for recognition of living foraminifera. *Contributions from the Cushman Foundation for Foraminiferal Research* 3, 56–60.
- Watkins, J.G., 1961. Foraminiferal ecology around the Orange County, California, ocean sewer outfall. *Micropaleontology* 7, 199–206.
- Word, J.Q., 1979. The infaunal trophic index. In: Segundo, E. (Ed.), *Annual Report 1978. Southern California Coastal Water Research Project*, Los Angeles, CA, pp. 19–41.